



UWL REPOSITORY

repository.uwl.ac.uk

An assessment of technological development and applications of decentralized water reuse: a critical review and conceptual framework

Maniam, Geetha, Zakaria, Nur, Leo, Choe, Vassilev, Vasilena, Blay, Karen and Behzadian, Kourosh
ORCID logoORCID: <https://orcid.org/0000-0002-1459-8408> (2022) An assessment of technological development and applications of decentralized water reuse: a critical review and conceptual framework. Wiley Interdisciplinary Reviews: Water, 9 (3).

<http://dx.doi.org/10.1002/wat2.1588>

This is the Accepted Version of the final output.

UWL repository link: <https://repository.uwl.ac.uk/id/eprint/8835/>

Alternative formats: If you require this document in an alternative format, please contact:
open.research@uwl.ac.uk

Copyright:

Copyright and moral rights for the publications made accessible in the public portal are retained by the authors and/or other copyright owners and it is a condition of accessing publications that users recognise and abide by the legal requirements associated with these rights.

Take down policy: If you believe that this document breaches copyright, please contact us at open.research@uwl.ac.uk providing details, and we will remove access to the work immediately and investigate your claim.



WILEY

An assessment of technological development and applications of decentralised water reuse - a critical review and conceptual framework

Journal:	<i>WIREs Water</i>
Manuscript ID	WATER-771.R1
Wiley - Manuscript type:	Overview
Date Submitted by the Author:	28-Jan-2022
Complete List of Authors:	Maniam, Geetha ; School of Pharmacy Zakaria, Nur; School of Chemical Engineering, Engineering Campus Leo, Choe; School of Chemical Engineering, Engineering Campus; Penang Green, the Institution of Engineers, Malaysia (Penang Branch) Vassilev, Vasilena; P+ Studio and Consulting Blay, Karen; School of Architecture Building and Civil Engineering Behzadian, Kourosh; School of Computing and Engineering Poh, Phaik; School of Engineering
Choose 1-3 topics to categorize your article:	Planning Water (WAAD) < Engineering Water (WAAA), Conservation, Management, and Awareness (WDAD) < Water and Life (WDAA), Water Quality (WCAD) < Science of Water (WCAA)
Keywords:	Decentralised, greywater, water reuse, rainwater, urban

SCHOLARONE™
Manuscripts

<p>London, UK</p> <p>k.behzadian-moghadam@exeter.ac.uk</p> <p>ORCID iD: 0000-0002-1459-8408</p>
<p>Poh Phaik Eong*</p> <p>School of Engineering, Monash University Malaysia, Jalan Lagoon Selatan, 47500</p> <p>Bandar Sunway, Selangor D.E., Malaysia</p> <p>poh.phaik.eong@monash.edu</p> <p>ORCID iD: 0000-0002-4215-5284</p>

Conflict of Interest

The authors declare no conflict of interest.

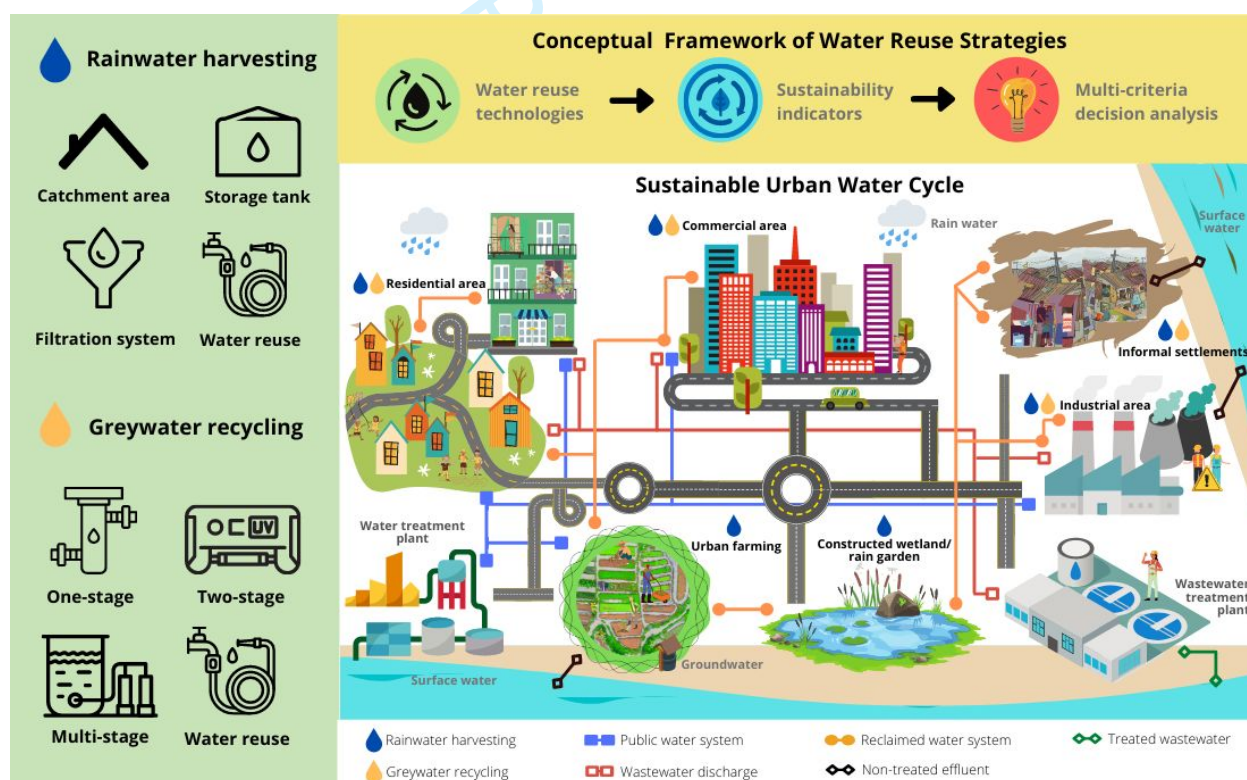
Abstract

Rapid development, urbanisation and population growth have contributed to water stress in many parts of the world. As freshwater sources are finite, it is essential to source for alternatives to ensure water security. Reclamation of non-conventional water sources (i.e.: rainwater and greywater) can be a viable alternative to alleviate water stress. In this work, various laboratory, pre-commercialisation scale and commercialised products for rainwater harvesting and greywater treatment and reuse were reviewed. As a result, a conceptual framework is proposed to provide an overview on the applicability of technologies under various settings, which were mapped against the intended use of treated water

and the potential for water supply expansion. This conceptual framework could aid decision makers in deciding on a suitable decentralised solution for water reclamation depending on limiting criteria. Decentralised systems of rainwater harvesting and greywater treatment for reuse are going to be crucial in reducing the dependence on the centralised water supply.

Keywords: Decentralised, greywater, rainwater, water reuse, urban

Graphical/Visual Abstract



Caption

Decentralised rainwater harvesting and greywater recycling systems are expected to play an important role to reduce the dependence on the centralised urban water supply and contribute to

1
2
3 water supply for urban farming and even informal settlements. Conceptual framework is suggested for
4
5
6 water reuse decision making.
7
8

9 **1. INTRODUCTION**
10

11
12 The world's main water sources consist of both saltwater and freshwater. Salt water contributes
13
14
15 97.5% of the world's water. Meanwhile, freshwater is made up of groundwater and surface water,
16
17
18 merely 2.5% of the world's water (UNESCO World Water Assessment Programme [UNESCO
19
20
21 WWAP], 2020). Freshwater withdrawals are used for agricultural, industrial or domestic uses.
22
23
24 Numerically, there is sufficient volume of freshwater for everyone on Earth. However, irregular water
25
26
27 distribution, wastage, pollution and poor water management add to the ever-increasing water tension.
28
29
30 Furthermore, population growth, rapid urbanisation and industrialisation with increasing irrigated
31
32
33 agriculture to meet the demand of staple food production contributed to increasing water demand.
34
35
36 The current population of 7.7 billion people is projected to grow by 26% in 2050 (UN, Department of
37
38
39 Economic and Social Affairs, Population Division, 2019). In the last century, water consumption has
40
41
42 escalated at more than twice the rate of population growth (UN, 2021). By 2030, it is estimated water
43
44
45 scarcity could displace 700 million people (UN, 2020). This situation is compounded by two factors;
46
47
48 water pollution and climate change.
49
50
51

52
53 Almost 2.3 billion people currently live in high water stress facing countries (UN-Water, 2021). The
54
55
56 majority of these countries from Sub-Saharan African, Asia and Oceania suffer from either physical
57
58
59 water scarcity or economical water scarcity reflected in the categories of high to extremely high stress
60

zones. Physical scarcity happens when there is insufficient water to meet demands, whereas economic scarcity occurs when there is lack of investment in water due mainly to water poverty. The SDG report of the UN-Water (2021) reveals that Western Asia and Northern Africa and Central and Southern Asia are withdrawing more than 70% of available water resources. The water scarcity problem can be addressed through various pragmatic ways such as water allocation to a variety of uses, reform of water management, changes in water policy and appropriate investments in order to ensure sustainability in water use. Innovative strategies have been elucidated to address challenges associated with water scarcity, which include increased funding in water-use efficiency, wastewater treatment, reuse and protection of water ecosystems (UN-Water, 2021). For example, water-use efficiency has increased by 9% globally since 2015 (UN-Water, 2021). Some strategic efforts to mobilise the water-use efficiency plan include water pricing policy, adding water metres to make consumers reduce their water usage, encouraging the use of household water efficient and water reuse appliances and fittings, to name a few.

One of the most efficient adaptation strategies for water supply and demand management is to use non-traditional water resources for water supply such as rainwater harvesting, greywater recycling and reuse. This is further supported by the fact that a high percentage of the consumed water worldwide is not treated, and the abundantly available annual precipitation is not tapped effectively. This situation must change due to the limited supply of safe water currently available, thus water reclamation is a progressive measure. This is in tandem with Target 6.3 of SDG 6 which addressed

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

the need to halve the proportion of untreated wastewater and substantially increase recycling and safe reuse globally. One of the indicators for achieving this target is to safely treat the proportion of domestic and industrial wastewater flow. There are various technologies for water reclamation and reuse throughout the world, especially centralised systems that follow existing international, national and local guidelines and regulations, However, recent advancement in water recycling technologies such as decentralised systems and green technology together with the emergence of complicated situations due to urbanisation like heavy pollution of freshwater sources and an increasing number of informal settlements in huge metropolitan cities, shift the research priorities toward the need to review the technological development and applications of water reuse to update water reclamation guidelines for small-scale technologies and form a conceptual framework for proper applications in urban areas (Angelakis et al., 2018; Reynaert et al., 2021).

Development of decentralised water recycling and reuse, especially in developing countries facing physical and economical water scarcity, could be a sustainable solution to provide safe, acceptable and affordable water to people from these localities. However, when deciding an appropriate system for implementation, many factors need to be considered such as the availability of technology, location of implementation and possibility of expansion. At present, there are limited sources of work that have reviewed the technological advancement for decentralised water recycling systems and used the conceptual framework to assess the applications of these technologies in different situations. Therefore, this paper aims to review domestic water recycling and safe reuse globally with a particular

focus on rainwater and greywater technologies together with the assimilation of water efficiency technological development in a conceptual framework for urban scale water reuse. The review presents crucial evidence for the scientific community and the stakeholders for adopting and developing water-reuse related strategies.

2. WATER RESOURCES AND CONSUMPTIONS

2.1 Traditional water sources

Freshwater is made up of surface water, groundwater, glacier and permafrost (UNESCO WWAP, 2020). Freshwater is the traditional main source of potable water. Potable water which is synonymous to drinking water is used in daily life for various reasons in different countries including drinking, cooking, basic hygiene, sanitation, garden watering, in addition to recreational, agricultural and industrial activities.

In developed countries such as the United States and the United Kingdom, on average 68% of community water system users receive water from surface water, while the rest receive it from groundwater (U.S. Department of Health & Human Services, 2009; Water UK, 2020). However, most of the tap water used in Australia comes from rainwater. Rainwater tanks are widely used to collect drinking water for rural areas in Australia (Environmental Health, 2020). The use of the rapid, low-cost provision of good quality groundwater as a potable water source is more evident in developing

countries or rural areas of developed countries across colossal areas of Asia, Africa and Latin America (Foster et al., 2000; Carrard et al., 2019). For example, Southeast Asia reported high rates of rural household groundwater consumption including Indonesia (90%), Timor-Leste (81%) and Myanmar (78%) (Carrard et al., 2019). In India, surface water and groundwater are utilised at around 690 km³ and 396 km³ respectively, annually (Kumar et al., 2005). It is estimated that 75% of the African population uses groundwater as the main source of drinking water (UN Water/Africa, 2004). Despite having river basins covering 64% of the continent's land area in Africa, only 22 - 34% of the population have access to safe water in at least 8 Sub-Saharan countries and in 2015, the population without access to potable water increased to almost half of global population (Tatlock, 2006; Ritchie & Roser, 2021a). The main reasons for this are insufficient infrastructure to mobilise the water source and transboundary conflicts (Musingafi, 2013). There is causality between economic development and water accessibility, chiefly from investment in water resources perspective (Dounghmanee, 2016).

Water withdrawals are defined as the permanent or temporary removal of freshwater from surface or groundwater sources to be channelled for purposes such as domestic supply, agriculture and industries (Ritchie & Roser, 2017). While the benefit of the groundwater outlined above, over-abstraction of groundwater through drilling and pumping has negative impacts on hydrogeology (Carrard et al., 2019). Excessive abstraction of groundwater will exhaust the aquifer over time and consequently disturb the hydraulic balance (Gun et al., 2021). This can also allow the intrusion of natural underground contaminants (arsenic), low-quality surface water, seawater or even sewage

leach from nearby areas, reducing the quality of groundwater and causing health effects (Gun et al., 2021). On the other hand, in many other places, surface water is commonly withdrawn for supply due to its easy accessibility. In some cases, river water can be collected by building dams and funnelled to treatment plants via a pipeline network. Dams are also built for other purposes than water supply such as flood control and, hydropower generation. Despite multiple benefits of having dams, they can have negative environmental, social and ecological impacts. Forced resettlement is faced by many vulnerable communities, mainly Indigenous communities who have a special bond with nature. For example, about ten thousand local people along the Bakui River in Borneo, Malaysia were relocated to a place 40 km downstream their customary land due to the construction of Bakun Dam, the largest dam in Southeast Asia (Aiken & Leigh, 2015; Cooke et al., 2017). They were also affected by the lack of social impact assessment of the megaproject development (Cooke et al., 2017). The project also destroyed the ecosystem by denuding the forest and disturbing fish migration (Aiken & Leigh, 2015). There is a strong need for governments and developers to mitigate the environmental and social effects of functioning dams. They need to critically evaluate the advantages and disadvantages of the construction of large dams for water withdrawals.

The current water withdrawal is escalating at a rate about 1% annually and the global water demand is estimated to increase by 20-30% by 2050 (UNESCO, 2021). The water withdrawal pattern of each country is strongly dependent on its socioeconomic construction. Agriculture contributes to 70% of freshwater withdrawal globally (Ritchie & Roser, 2017). In 2010, India was the largest agricultural water consumer, at nearly 688 billion m³, followed by China at 388 billion m³ and the United States at about 175 billion m³ (Ritchie & Roser, 2017). Agriculture, for many developing countries, contributes a smaller share of income compared to other sectors, however, consumes the largest proportion of

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

water, where more than 90% water resources of developing countries are allotted to agriculture (Doungmanee, 2016; UNESCO WWAP, 2017). For industrial water uses, the United States is the biggest consumer at nearly 250 billion m³, almost double that of China (Ritchie & Roser, 2017). In general, it can be observed that countries across Europe, Americas, Pacific and East Asia consume more than a billion m³ annually and the rates for Sub-Saharan Africa and South Asia reported at the low side accounting an average volume of 500 million m³ (Ritchie & Roser, 2017). For higher-income countries, the industrial sector forms a bigger share of total gross domestic product and provides a larger allotment of employment, and thus, contributing to a higher industrial water withdrawal share. In terms of municipal water withdrawal, China with the highest population number dominated the chart, followed by the United States and India (Ritchie & Roser, 2017). Even though the United States had an almost 4 times smaller population than India, it had higher domestic water consumption per capita. In many nations, especially in developing countries, population growth and urbanisation intensify the pressure on many water resources, causing rapid depletion of supply (Ahuja, 2019). In general, while the strategy of investing in water infrastructure to expand the network water supply by tapping freshwater sources is necessary to bring on modernisation and upgradation of water delivery systems (Rosegrant et al., 2002; Peter-Varbanets et al., 2009), improving the efficiency of water use is the most cost-effective and immediate way to conserve water for water sustainability.

2.2 Reclamation of non-conventional water sources

Meeting growing demand by developing new water supplies has only limited potential and focus is increasingly being placed on exploring non-conventional water sources. Figure 1 illustrates the non-drinking water sources available from a dwelling compound which include greywater, blackwater, groundwater and rainwater. Greywater originates from the bathroom sinks, showers and washing machines, while blackwater comes from toilets and kitchen sinks. Blackwater is often composed of high nutrient-load and pathogens from faecal matter and food waste, in addition to oil-waste from cooking. These microbes are disease-causing. In contrast, greywater contains 80% less organic matter, suspended solid and nutrient-load than blackwater (Hocaoglu et al, 2010). Blackwater together with greywater constitute the domestic wastewater. Even though many countries recycle wastewater for non-potable uses, there are few countries including Singapore, Australia, Namibia and few states like California, Virginia and New Mexico that produce drinking water from reclaimed wastewater (Leong & Lebel, 2020). Recycling of domestic wastewater is more stringent due to inherent contaminants and the concept of 'toilet-to-tap' is a psychological barrier among consumers. Rainwater is another readily available water source with minimal treatment required as it is free from natural contaminants. On the other hand, desalination is the process of recycling abundantly available salt water. However, the process of salt water reclamation is very costly and requires use of large areas of land compared to other reclaimed water (Hinkebein, 2004). Recycled water that is not subjected to the same processing restriction as drinking water can be used as a non-drinking water system for many purposes such as watering plants, flushing toilets or washing house porches or cars. This will help in reducing the consumption of potable water from freshwater for non-potable usage.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

Rainwater harvesting and greywater recycling are the easily adaptable strategies for safe water reuse due to inherent low level of water contamination, lower cost and land required depending on the system design (Oh et al., 2018; Boano et al., 2020). Even though these innovations are gaining popularity, their implementation is limited due to regional differences such as water price, availability and technological accessibility (Peter-Varbanets et al., 2009).

For Peer Review

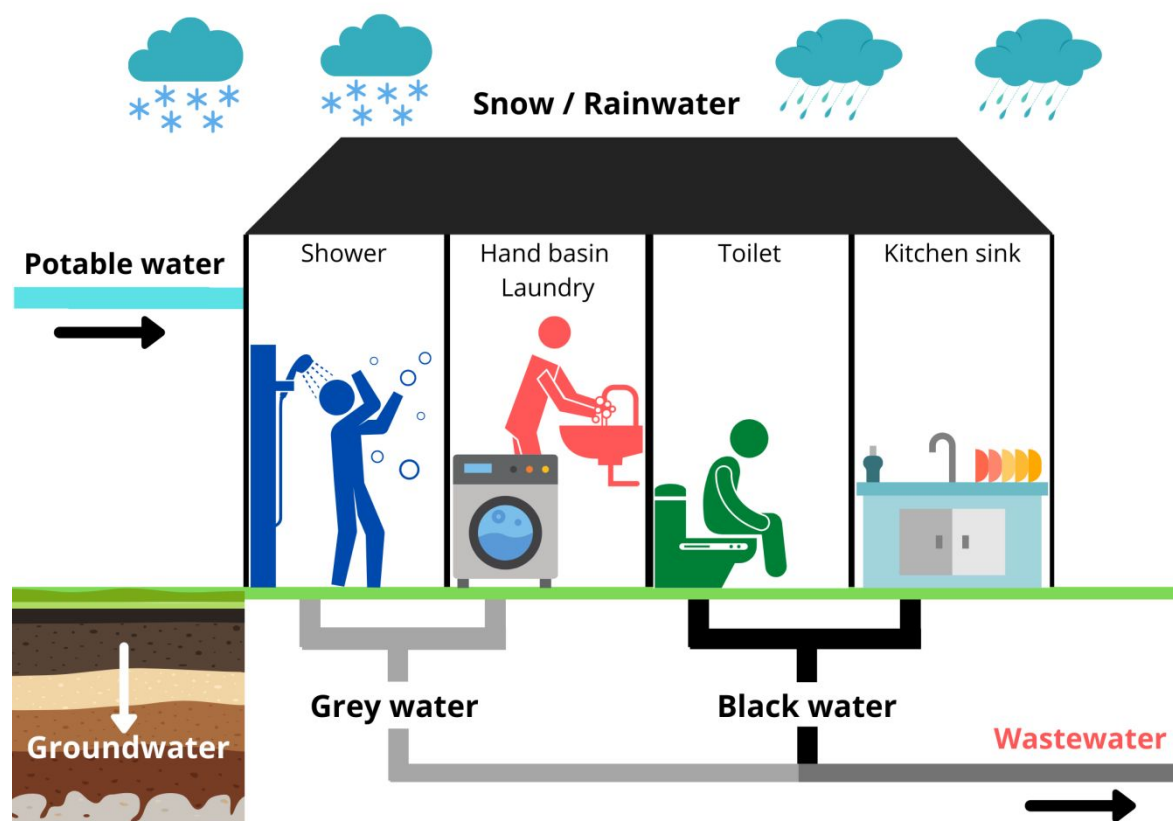


Figure 1. Schematic representation of non-drinking water sources available from a house compound which include greywater, blackwater, groundwater and precipitations.

2.2.1 Characteristics of rainwater and greywater

The quality of rainwater and greywater can be measured by physical, chemical and microbiological properties. Physical parameters are turbidity, total suspended solids (TSS) and total dissolved solids (TDS), whereas the chemical parameters are biological oxygen demand (BOD), chemical oxygen demand (COD), nitrates, phosphates, total nitrogen (TN), total phosphorus (TP), hardness and heavy metals. Enumeration of total coliform, faecal coliform and *Escherichia coli* (*E. coli*) are examples of microbiological quality standards. Rainwater in general is free from inherent impurities. However,

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

composition of rainwater collected is heavily affected by different catchment areas. Rainwater collected from land run-offs is generally more polluted than rooftop as reflected by the higher values of TSS, TDS, COD and bacterial enumeration; and atmospheric rainwater is the least polluted (Table 1). Level of heavy metals leached into collected rainwater varies according to the type of roof coating materials. Other factors affecting rainwater quality include local emissions, climate conditions, faecal dropping from animals or even heavy metals and impurities leached from the catchment area due to the slightly acidic nature of rainwater. Even though annual rainfall in tropical climate regions averaged at 59 inches, unpredictable rainfall in certain parts of the world with dry, mild, continental and polar climate might limit dependence on rainwater as a standalone water source (National Geographic, 2017). Contrariwise, the amount and quality of greywater produced by each household varies and significant differences can be observed between urban and rural regions in a country and across the countries globally (Bodnar et al., 2014). The amount of greywater produced vary from as low as 14 L/day in regions such as Africa and Middle East to as high as 225 L/day per person in other parts of the world (Morel & Diener, 2006; Halalsheh et al., 2008; Al-Hamaiedeh & Bino, 2010). The generation of greywater is also dependent on various factors including number of household occupants, water supply availability, lifestyle, geographical location, water, sanitation and hygiene infrastructures and climate (Oteng-Peprah et al., 2018). Jamrah et al. (2006) reported 69% of domestic water consumption attributable to greywater production.

Table 1. Physical, chemical and biological properties of rainwater at different catchment areas. Abbreviations: TSS- total suspended solids, TDS- total dissolved solids, COD- chemical oxygen demand.

Rainwater quality parameters	Catchment areas		
	Atmospheric	Rooftop with different coating materials	Storm water
pH	4.4 - 8.7	4.6 - 8.1	6.4 - 8.3

Electrical conductivity (dS/m)	0.0038 - 0.0448	0.43 - 2.5	2.65
Alkalinity (mg/L)	4 - 43	54 - 150	132 - 651
Hardness (mg/L)	0 - 71	5 - 88	30 - 1305
Turbidity (NTU)	-	0.1 - 22.3	1.4 - 47.6
TSS (mg/L)	4 - 211	20 - 460	40 - 866
TDS (mg/L)	23 - 114	26 - 404	87 - 1693
Nitrates (mg/L)	0.018 - 3.60	0.1 - 17.0	4 - 120
Phosphates (mg/L)	0.01 - 0.04	0.03 - 0.5	0.1 - 4.03
Lead (mg/L)	0 - 0.004	0.001 - 0.01	0.028
Iron (mg/L)	0 - 0.008	0 - 0.19	0.038 - 4.96
Zinc (mg/L)	0.0012 - 0.02	0.03 - 1.46	0.085
COD (mg/L)	-	23 - 43	33 - 279
Faecal coliform (MPN or CFU*/ 100 mL)	0*	3	11
<i>E. coli</i> (MPN or CFU*/ 100 mL)	0*	2	5 - 507
References	(Radaideh et al., 2009; Samuel et al., 2012; Chukwuma et al, 2013; Abulude et al., 2018; Liyandeniya et al., 2020) (Radaideh et al., 2009; Samuel et al., 2012; Lee McCarthy et al., 2012; et al., 2017; Khayan et Kayhanian et al., 2012; al., 2019; Abuefutoh et Samuel et al., 2012; al., 2020) Salehi et al., 2020)		

The greywater makes up about 75% of the wastewater volume in households (Leal et al., 2010; Edwin et al., 2014) and the high percentage has led many countries to adopt greywater recycling as a

supplementary source to freshwater supply. Quality of greywater also varies according to sources of origin such as shower, laundry and washbasins, in addition to factors addressed for variation in the quantity of greywater generated (Oteng-Peprah et al., 2018). Physical, chemical and biological properties of greywater from different regions are summarised in Table 2. The Asian and African households have high turbidity and TSS in their greywater. Greywater generating from kitchen and laundry generally have high turbidity and suspended solids due to actions of washing vegetables or clothes that contribute to food particles, oils, sand, clay and fabric fibres (Shaikh & Ahammed, 2020; Boano et al., 2020). In addition, the American and Asian households were reported to have high TN compared to other counterparts however within the range reported in literature (4 - 74 mg/L) (Oteng-Peprah et al., 2018). While high level of nitrogen in greywater originates from kitchen waste, elevated nitrates and phosphates level originate from the use of soap, laundry detergents and dishwasher (Boyjoo et al., 2013). Another interesting finding was that the European households showed highest BOD (208 - 1363 mg/L) and COD (390 - 2072 mg/L) ranges in their greywater, followed by the Middle-East. BOD measures the biodegradable organic load in water, while COD measurement include both organic and inorganic matters susceptible to oxidation. The BOD/COD ratios ranging from 0.31 to 0.71 indicate almost 50% of the organic compounds in the greywater are biodegradable (Halalsheh et al., 2008; Oteng-Peprah et al., 2018). The elevated level of BOD and COD in water conventionally due to the presence of emerging pollutants, the xenobiotic organic compounds such as pharmaceuticals, personal care products, endocrine-disrupting compounds etc. (Oteng-Peprah et al., 2018; Shaikh & Ahammed, 2020). The total coliforms and *E. coli* counts were almost similar across

the regions, except low total coliforms observed for the Middle-East. Bacteria in greywater could be from the kitchen, hand basin or bathroom sources.

Different greywater characteristics across the different continents implied that there would not be a 'one size fits all' solution to reclaim greywater for reuse but dependent on the localised characteristics of the greywater and intended use of the treated water. Other factors that could be decisive on the method to implement rainwater harvesting or greywater treatment and reuse are (i) technology readiness, (ii) cost to implement and maintain the system and (iii) usability of the reclaimed water, which will be further discussed in subsequent sections.

Table 2. Physical, chemical and biological properties of greywater globally. Abbreviations: TSS- total suspended solids, TN- total nitrogen, TP- total phosphorus, BOD- biological oxygen demand, COD- chemical oxygen demand.

Greywater quality parameters	Countries					
	Americas	Oceania	Europe	Middle-East	Asia	Africa
Turbidity (NTU or °FTU)	34 - 100	-	93*	15 - 58	196 - 225	270
TSS (mg/L)	19 - 156	0.5-795	11 - 319	45 - 155	55 - 619	180 - 537
Nitrates (mg/L)	0.11	-	-	0.70	1.49 - 40	4.26 - 8.43
TN (mg/L)	4.3 - 50.3	-	0.5 - 15.0	10.6 - 24.1	29.7 - 41.2	17
TP (mg/L)	0.5 - 5.3	-	0.1 - 187	1.4 - 12	1.0 - 5	2.3 - 61
BOD (mg/L)	45 - 144	3 - 380	208 - 1363	335 - 568	5 - 445	204 - 350

COD (mg/L)	205 - 600	-	390 - 2072	380 - 1171	89 - 643	644 - 848
Total coliforms (CFU or MPN*/100 mL)	4.0x10 ⁵ - 7.6x10 ⁶	4.4x10 ² - 3.7x10 ⁸	-	0 - 2.3x10 ¹ *	-	3.8x10 ⁶
<i>E. coli</i> (CFU/100 mL)	7.0x10 ² - 5.06x10 ⁴	5.0 - 8.6x10 ³	-	1.1x10 ⁴	-	1.8 - 2.4x10 ⁴
References	(Zavala et al., 2016; Chrism & Nolasco, 2017; Palmarin & Young, 2019; Goncalves et al., 2021) (Leonard et al., 2016) (Noutsopoulos et al., 2018; Sievers & Londong, 2018; Jabri et al., 2019; Truu et al., 2019) (Abdel-Shafy et al., 2019; Oktor & Celik, 2019; Alrousan & Dunlop, 2020; Uceqli & Kaya, 2021) (Deng et al., 2020; 2020; r t al., 2020; Perdana et al., 2020; Subramaniam et al., 2020) (Oteng-Pepurah et al., 2018; Dwumfour-Asare et al., 2020; Raphael et al., 2020)					

2.3 Decentralised systems

Water efficiency entails encouraging the sustainable use of water including implementing measures that allow for significant reductions in water wastage, pollution and environmental impacts in the production of commodities and services including domestic water provision (UNEP, 2014). Water efficiency is a practical way to reduce the consumption of the depleting freshwater sources globally. Water efficiency is applicable to both potable water and non-drinking water. Non-drinking water includes rainwater, domestic wastewater and urban (domestic + industrial) wastewater. Wastewater (also called sanitary sewage) normally consists of faecal coliforms (mainly from domestic areas), hazardous inorganic chemicals (heavy metals, nitrates), organic pollutants (pesticides, fertilisers) and emerging contaminants (pharmaceuticals, microplastics). Wastewater can cause health burdens to

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

mankind and disturbs natural aquatic life cycle (Tortajada, 2020). Thus, wastewater treatment is necessary to process the wastewater to meet the environmental discharge standards or other quality standards for purposes such as recycling or reuse (Tortajada, 2020; UN, 2020). Besides being the predominant environmental buffer of wastewater discharge, surface water especially rivers is the main source for providing raw water for drinking water treatment plants and hence acting as direct drinking water source for many communities, especially in remote locations of developing nations (Tortajada, 2020). Globally, 1.56% of the population still rely on surface water as drinking water (Ritchie & Roser, 2021). The number rocketed to 6.84% focusing on just the Sub-Saharan African population (Ritchie & Roser, 2021). Large change in feed chemistry due to untreated wastewater will affect the water treatment efficiency in the treatment plant and could lead to plant malfunction affecting the drinking water supply of even a larger population. These stress the need to comply with the wastewater discharge standards and regulations.

However in real scenarios, non-compliance to wastewater discharge standards is observed (OECD/Eurostat, 2018). Globally, around 80% of sewage is discharged without prior treatment or recycling. The high-income countries on average treat 70% of the generated wastewater, followed by upper-middle-income countries (38%), lower-middle-income countries (28%), and low-income countries, where only 8% of the wastewater generated is treated (Sato et al., 2013; UNESCO WWAP, 2017). Based on SDG 6 Summary Progress Update 2021, out of 128 countries' data, 56% of the household wastewater is safely treated (UN-Water, 2021). However, no or very little data is available

on the proportion of the industrial and agricultural wastewater flow that is safely treated. Untreated wastewater could possibly be explained by the lack of connection between wastewater sources and the centralised wastewater treatment plant. For example, in Poland, 35% of the rural population are not connected to sewage systems (Boguniewicz-Zablocka & Capodaglio, 2017). This is more commonly observed in developing countries. Possible reasons for the lack of centralised wastewater treatment facilities in the rural areas of developing countries are, (i) situated further from major treatment centres, (ii) huge investment required for centralised systems in remote locations and (iii) lack of financial resources and funding support (Peter-Varbanets et al., 2009; Boguniewicz-Zablocka & Capodaglio, 2017). The centralised wastewater treatment systems consist of collection, treatment, disposal, collection and discharge or distribution (if treated water is reused). The sewage collection systems also need an extensive piping infrastructure (Estévez et al., 2021). In addition, the complex and scattered topography of rural areas would increase the capital and maintenance cost of the pipeline (Estévez et al., 2021; Huang et al., 2021). In 2018, only a quarter of China's rural municipal wastewater was adequately treated, where insufficient fund was one of the contributing factors of the treatment plant infeasibility (Huang et al., 2021). Globally, some countries experience a funding gap of 61% to achieve the SDG water and sanitation targets and only less than 15% of nations have the financial capacity to execute their water development plans (UN, 2019; UN, 2020).

In the next 30 years, the wastewater effluents are expected to rise up to 10 - 15% further contributing to water quality degradation (Ahuja, 2019). Wastewater is not only a source of contamination of

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

surface and groundwater resources, but the presence of emerging contaminants will also add complexity to the wastewater recycling process. This will increase the by-now high cost of treatment and management of reclaimed wastewater through centralised systems, making large centralised wastewater treatment systems not the most viable option in urban, peri-urban or in rural water management. One of the SDG 6 global targets explicitly stated that, by 2030, the need to “expand international cooperation and capacity-building support to developing countries in water and sanitation-related activities and programmes, including water harvesting, desalination, water efficiency, wastewater treatment, recycling and reuse technologies” (UN-Water, 2021). In line with this, localities who lack a centralised wastewater treatment infrastructure could benefit from implementing lower cost and safe water reuse options such as decentralised systems. Decentralised systems could be of a single household or a community scale system. UNESCO WWAP (2017) reported the investment, operation and maintenance costs of decentralised systems are lower than large-scale systems, in the range of 20 - 50% or even more. Thus, rainwater and greywater recycling schemes are feasible small-scale technologies for water reuse. However, there are multiple factors that might affect the sustainability and viability of developing decentralised systems especially in developing countries. The implementation of decentralised water reuse systems is particularly challenging in countries with low water tariff (Chirisa et al., 2017; Lani et al., 2018). In these cases, various strategies such as incentives, microcredit schemes and subsidies by relevant authorities or the government can be introduced to encourage the adoption of decentralised technologies, preceded with social, economic and ecological awareness of these water efficient systems. At the government

scale, changes in policy and regulation on the implementation of the decentralised schemes as part of integrated water resources management, especially in urban settings that are facing high water stress, are necessary to highlight their importance and for quality control.

2.3.1 Water quality requirements for various applications

Recycling and reclamation of rainwater and greywater produce water with different qualities depending on the level of water treatment applied. There are different standards adopted by different nations for various purposes such as garden irrigation, agricultural irrigation, toilet flushing, cooling process or even laundry. Table 3 shows a summary of some standards for main quality parameters in municipal and agricultural water reuse. As it can be seen in the table, reclaimed water for domestic uses has stricter water quality requirements than for agricultural ones, especially in terms of bacterial enumeration. In addition, standards in developed countries such as the United States and the European nations seem to be stricter for both municipal and irrigation water reuse than other countries such as China and Namibia. The water quality standards might be a reflection of a supportive policy environment. The US has written, "Guidelines for Water Reuse" produced in 2012 with each state having its own regulations for reclaimed water (EPA, 2012; Niekerk & Schneider, 2013). However, this may not be the case for other developing countries. In Namibia, currently there is no established law on water reuse, but "The Water Supply and Sanitation Policy" is in place to educate its citizen on water recycling and reclamation (Niekerk & Schneider, 2013).

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

It is essential for countries to establish guidelines or standards for water reclamation as it allows proper studies to be carried out on different scales of technology and to optimise the process to produce treated water that is safe for consumption, particularly if decentralised treatment systems were to be set up to alleviate water stress. The users will most likely not have expert’s knowledge in water treatment and reclamation. When clear standards have been developed, proper measurements can be taken and control measures can be implemented in decentralised treatment systems for compliance and protect users from water-borne diseases.

For Peer Review

Table 3. Reclaimed water quality standards for municipal and agricultural water reuse. Abbreviations: TSS- total suspended solids, BOD- biological oxygen demand, COD- chemical oxygen demand.

Reclaimed water quality parameters	Municipal water reuse			Agricultural water reuse			
	United States	China Reclaimed	Namibian Water Act 1956*	United States	European	China Reclaimed	Namibian Water Act 1956**
	Environmental	Water Reuse		Environmental	Commission- Joint	Water Reuse	
	Protection Agency	Regulations GB/T		Protection Agency	Research Centre	Regulations GB	
	2012	18920-2002		2012	2017	20922-2007	
pH	6.0 - 9.0	6.0 - 9.0	6.0 - 9.5	6.0 - 9.0	-	5.5 - 8.5	4 - 11
Turbidity (NTU)	≤ 2	≤ 5	1 - 5	≤ 2	≤ 5	-	10
TSS*/ TDS** (mg/L)	-	≤ 1500**	-	-	≤ 10*	60 -100*	-
Residual chlorine (mg/L Cl ₂)	1.0	1.0	0.1 - 5	1.0	-	1.0 - 1.5	0.1 - 5
Ammonia (mg/L N)	-	≤ 10	1 - 2	-	-	-	4
BOD (mg/L)	≤ 10	≤ 10	-	≤ 10	≤ 10	40 - 100	-
COD (mg/L)	-	-	-	-	-	100 - 200	-
Total*/ Faecal**/ <i>Escherichia</i> *** coliforms (CFU/ 100 mL)	0 **	≤ 3 *	0 - 10 * 0 - 5 ** 0 ***	0 **	≤ 10 or below detection limit ***	20000 a/L **	100 * 50 ** 10 ***

References	(EPA, 2012)	(Rodrigues, 2021)	(Niekerk & Schneider, 2013)	(EPA, 2012)	(Alcalde-Sanz & Gawlik, 2017)	(Rodrigues, 2021)	(Niekerk & Schneider, 2013)
------------	-------------	-------------------	-----------------------------	-------------	-------------------------------	-------------------	-----------------------------

*Group A- Water with an excellent quality, Group B- Water with acceptable quality; **Group C- Water with low health risk, Group D- Water with a high health risk

For Peer Review

3. LABORATORY SCALE AND PRE-COMMERCIALISED WATER TREATMENT

SOLUTIONS

While there are existing solutions in the market, it is imperative that laboratory scale and pre-commercialised water treatment solutions are reviewed and explored as they could potentially address the world's growing demand for clean and affordable water. The solutions reviewed will be categorised according to their capabilities in primary, secondary and tertiary treatment of water.

3.1 Primary treatment methods

Coagulation-flocculation is conventionally effective in removing suspended solids and organic matters from water and wastewater. Examples of coagulants and flocculants are ferric chloride and alum. Ferric chloride at a dosage of 60 mg/L could remove 59% of COD and 73% of TSS from greywater (Antonopoulou et al., 2013). Although ferric chloride works in a wide range of pH, it is very corrosive in nature. On the other hand, alum ($\text{Al}_2(\text{SO}_4)_3 \cdot 18\text{H}_2\text{O}$) removed up to 83% of TSS and 99% of *E. coli* in greywater at pH 7.5 (Ghaitidak & Yadav, 2015). Coagulants and flocculants are rarely used as a single-treatment option and commonly paired with other treatment methods such as sedimentation, sand filtration or disinfection. Furthermore, the flocs from coagulation are secondary wastes that could pose environmental issues if not dealt with properly. Therefore, there is a constant search for improved coagulants, especially on the use of coagulants derived from natural sources (i.e. chitosan, *Moringa oleifera*, diatomite) and laboratory scale studies were conducted to identify the optimal coagulant dose for various applications. Diatomite (1.0 - 2.5 mg/L) added to 50 NTU turbid water

during sedimentation stage produced 0.87 log reduction of *E. coli* at pH 5 (Sha'arani et al., 2019).

While the laboratory scale studies have shown promising results, the fact that need for conducting detailed study to identify an optimal coagulant dosage has limited the deployment of new types of coagulants for the market due to the need for involvement of an expert in the field rather than a technology managed by non-experts.

Electrocoagulation was also studied to remove organic and inorganic compounds in greywater in the laboratory scale. The electrocoagulation systems consist of a cathode which hydrolyses water into hydrogen gas and hydroxyl groups, as well as an anode which introduces metal ions as shown in Figure 2. It is advantageous to work with electrocoagulation for contaminants removal in water without need for the addition of chemicals as metallic ions as coagulants are continuously generated from the electrode when current is applied. Only the sacrificial anode, frequently made of aluminium, iron and magnesium has to be replaced periodically. In recent years, electrocoagulation process has been integrated with different treatment processes such as membrane filtration (Khosravanipour et al., 2019), ozonation (Barzegar et al., 2019), and electro-disinfection (Cotillas et al., 2020) to achieve higher water treatment effectiveness and efficiency. The implementation of electrocoagulation meanwhile has been impeded since it needs a number of issues to be solved at a larger scale namely: (i) treatment of sludge produced from the process; (ii) power requirements that would make it difficult to deploy in areas with energy security issues, for example: in rural communities without consistent electricity supply; and (iii) lack of sufficient manpower and experts to maintain the system.

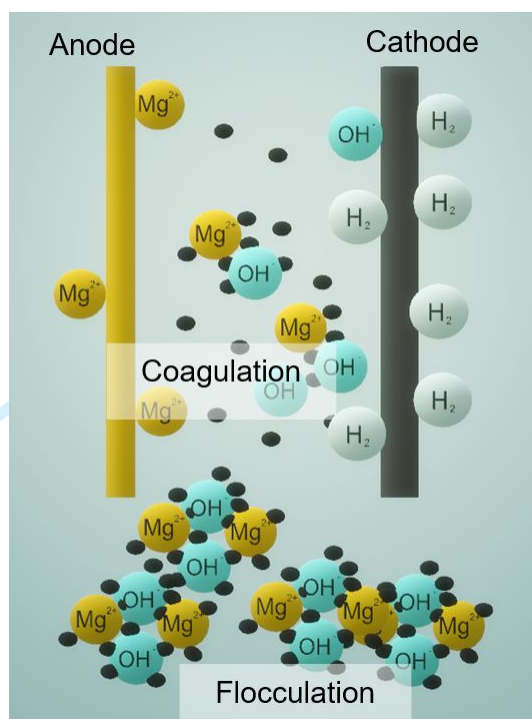


Figure 2. The coagulation and flocculation processes in electrocoagulation, where in the cathode, water hydrolysed into hydrogen gas (H_2) and hydroxyl groups (OH^-), while in the anode, metal ions (Mg^{2+}) are produced.

Membrane filtration is also often explored in the laboratory scale for water purification given the ease of use. Manouchehri and Kargari (2017) studied the microfiltration of laundry effluent and found that

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

COD (2538 mg O₂/L) and BOD (1190 mg O₂/L) was reduced more than 73.4%, however, the values were higher than water reuse standards. Ultrafiltration could reduce the turbidity and total suspended solids of treated wastewater from a conventional sewage treatment plant up to less than 0.2 NTU, besides lowering COD in the range of 20-60 mg/L (Falsanisi et al., 2010). Hourlier et al. (2010) reported commercial, PCI nanofiltration membranes reduced COD and anionic surfactant in greywater up to 95%, however a very high pressure is required to achieve reasonable water flux. While various membrane materials showed good performance in the treatment of greywater, they have not been widely used for commercial purposes, the difficulties for producing the membranes in huge quantities. Furthermore, the energy requirement to operate a membrane filtration module is high, making it less accessible in areas without stable electricity supply, or for lower income communities. Therefore, the suitability of using membrane filtration for decentralised water reuse remains a challenging question that depends on several factors as outlined above plus capital investment and operational costs.

3.2 Secondary treatment methods

Biological treatment is often employed as a secondary treatment process to remove biodegradable organic and suspended solids in any types of wastewaters. Centralised treatment systems utilise activated sludge to remove contaminants in wastewater aerobically. Unlike the centralised wastewater treatment systems, the decentralised systems use other biological treatments including biofilter, rotating biological contactor and membrane bioreactor that have smaller footprint. Biological treatment

1
2
3 systems are well studied and are used commercially. However, research works are still being
4
5
6 conducted to improve the efficiency of the system, or to reduce the time required to treat the
7
8
9 wastewater.

10
11
12
13
14
15 Two types of biofilters are available for greywater treatment, namely macrobiofilters and membrane
16
17
18 biofilters. The microorganisms are either attached or suspended in macroporous filters, while
19
20
21 membrane biofilters can be submerged into wastewater or installed at the side stream. Santos and
22
23
24 Daniel (2020) reviewed the activated carbon filters developed for water and wastewater treatment.
25
26
27 The biofilm formation and activity are influenced by the feed water quality, backwashing, hydraulic
28
29
30 conditions, temperature. The activated carbon biofilters could remove 17 - 63% of organic matter from
31
32
33 wastewater (Hoang et al., 2004; Rhim, 2006; Hoang et al., 2008; Kalkan et al., 2011; Reungoat et al.,
34
35
36 2011; Reungoat et al., 2012; Pramanik et al., 2014). Sharaf and Liu (2021) treated greywater using
37
38
39 activated carbon biofilter recently and found that total COD removal near to 70% was achievable after
40
41
42 24 hours. Germec et al. (2020) also reported an in-sight summary on the biofilm reactors with value-
43
44
45 added products such as bioenergy, carbohydrates, antibiotics, acids, vitamins and enzymes.

46
47
48
49
50 Anaerobic biofilter was also integrated with an up-flow anaerobic sludge blanket to treat the black
51
52
53 water for simultaneous biogas production (Dorji et al., 2021). Anaerobic biofilter removed up to 98% of
54
55
56 COD with a hydraulic retention time of 8 hours. The moving bed biofilm reactors could reduce 59% of
57
58
59 BOD, 70% of COD, 87.07% of TSS, 66% of turbidity of greywater if paired with a settling tank
60

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

(Chrispim & Nolasco, 2017). At the full scale, the multi-stage moving bed biofilm reactors integrated with sand filter and ultraviolet (UV) disinfection reduced 94.5% of COD, 99.4% of BOD, 91% of DOC, and 99.7% of NH₄-N from the greywater (Saidi et al., 2017). Al-Wasify et al. (2018) treated greywater using a pottery scraps column, followed by a moving bed biofilm and H₂O₂ disinfection. The overall COD removal of 95%, TSS removal of 97.8%, BOD₅ removal of 95.1%, and turbidity removal of 93.6% were attained. Besides that, Eslami et al. (2017) utilised a fixed bed biofilm with activated sludge system to remove 92.52% of COD from greywater. This system also removed linear alkylbenzene sulfonate (LAS), oil and grease as much as 94.24% and 90.07%, respectively. Tombola et al. (2019) reported that the sequencing batch biofilm reactor alone could remove COD as much as 86.5%, ammonium up to 98.4% and TN near to 71%. COD removal up to 95% could also be achieved after 120 days using multi-chambered anaerobic biofilm reactors with *Proteobacteria* sp. and *Metahnogenic* sp. in the treatment of greywater (Khuntia et al., 2019). Zheng et al. (2020) commented that *Proteobacteria*, *Bacteroidetes*, *Pseudomonas*, *Enterobacter* in small diameter gravity sewers could attain COD removal up to 22.31%, TN removal at 21.95% and TP removal at 6.16%.

Pronk et al. (2019) used microfiltration and ultrafiltration membranes with biofilm for water and wastewater treatment. Jabornig and Podmirseg (2015) treated greywater using PVDF membrane (0.2 µm) with biofilm. The turbidity removal was 97%, COD removal was 69%, and TSS removal is 99%.

Zhou et al. (2020) recently applied a membrane biofilm reactor to greywater as well. The COD removal was improved to 88.4%, with LAS removal of 95.6% and TN removal of 80% were recorded.

Microfiltration or ultrafiltration membranes are installed in a membrane bioreactor with a small footprint to retain the solid materials produced from biological treatment. Cecconet et al. (2019) summarised the research works related to greywater treatment using membrane bioreactor. Membranes with pore size smaller than 0.45 μm were commonly used, while hydraulic retention time in the range of 0.3 to 40.8 hours was mostly reported. BOD less than 10 mg/L as stated by World Health Organization (WHO) standards could be achieved.

Although great results can be achieved with biological treatment systems in the treatment of greywater, the system is very sensitive to changes and requires to be constantly operated and monitored by experts which could be challenging for many local users. Hence for biological treatment systems to be successfully implemented at a decentralised level, the expertise gap needs to be addressed. In addition to the future development to streamline the expertise requirements, the introduction of artificial intelligence as means to control and monitor progress remotely can be a good alternative for resolving this challenge.

3.3 Tertiary treatment methods

Disinfection is essential given the treated wastewater will be reclaimed for reuse. The level of microorganisms in the effluent should be within an acceptable concentration to prevent health issues during reuse. Chlorine is a common disinfectant in drinking water, available in the form of compressed

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

gas, sodium hypochlorite solution or calcium hypochlorite, reacting with water to form hypochlorous acid and hydrochloric acid. The acid produced then kills microorganisms through the destruction of the cell wall and chemical bonds in the cell. However, the use of chlorine will produce harmful by-products such as THM which is cancerogenic for human health. As such, 66 countries have specified a guideline value for residual chlorine lower than the WHO guideline value of 5.0 mg/L (World Health Organization, 2018). Thus, there is a constant search in the laboratory to identify alternative disinfectants that are less harmful to human health and any other living organisms for possibility of reuse.

Performic acid with a concentration range of 0.5 - 1 mg/L was studied and was found to be able to deactivate S2-coliphages, DNA-coliphages, enterococci and *E. coli*, *Salmonella* spp., *Clostridium perfringens* spores and Giardia cysts in 10 minutes (Karpova et al., 2013). Peracetic acid (PAA) forms less halogenated disinfection by products compared to chlorine. Peracetic acid with a concentration in the range of 2-15 mg/L deactivated bacterial (*E. coli*, *Enterococcus spp.*) effectively within 30 min, but less effective than sodium hypochlorite in the disinfection of secondary wastewater at low temperature (4 °C) (Hassaballah et al., 2020). Given that acids are corrosive materials, it is not safe for implementation in a decentralised treatment system if the system is to be handled by individuals who are not familiar with the risks associated to the use of acids and proper methods to handle them.

Alternatively, solar disinfection can achieve a reduction in 2.47-log of total coliforms and 3-log of faecal coliforms to meet WHO standard in a study conducted using a tubular photoreactor (Igoud et al., 2015). UV dose of 10 mJ/cm² is sufficient to reduce *E. coli* to 5000 CFU/100 mL for discharge to surface water, but UV dose of 30 mJ/cm² is required to attain *E. coli* reduction to 10 CFU/100 mL for unrestricted reuse purposes (Antonelli et al., 2008). However, UV disinfection can be affected by turbidity, suspended solids concentration, colour, organic molecules, and UV absorbing components such as Fe²⁺ ions. Besides solar and UV-C (wavelength less than 280 nm) disinfection, UV-LED with long life span and low energy consumption has been studied in water disinfection (Chen et al., 2017). Moreover, photochemical and photocatalytic disinfection have been extensively studied in recent years. UV irradiation induces the formation of hydroxyl, chlorine, and sulphate radicals during photochemical disinfection in the presence of hydrogen peroxide (H₂O₂), chlorine, persulfate, and other chemicals (Zhang et al., 2019; Moreno-Andrés et al., 2020). Formisano et al. (2016) reported that sunlight/H₂O₂ and sunlight/PAA disinfection are effective to reduce *E. coli* in the wastewater to 3.52-log of *E. coli* and 3.13-log of *E. coli*, respectively in 120 min. On the other hand, UV irradiation also induces the formation of hydroxyl and oxygen radicals from water in the presence of photocatalysts such as TiO₂, ZnO, and more. Pablos et al. (2012) compared the effectiveness of TiO₂ free particles and TiO₂ fixed bed reactors in the deactivation of *E. coli*. TiO₂ free particles reduced *E. coli* under the detection limit in less than 160 min, while TiO₂ fixed bed reactor requires more than 320 min to reduce *E. coli* under the detection limit. However, what limits the implementation of such

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

systems are the space required, cost and accessibility to advanced materials for implementation in decentralised treatment system.

Ozone is a strong oxidant and virucide that can be produced by transferring the dry air or oxygen through a high voltage electric discharge. It oxidises or destroys the cell wall, damages the nucleic acid constituents, and depolymerizes through the breakage of C-N bonds. However, ozone is corrosive and corrosive- resistant materials should be used in the ozone treatment systems. Liberti et al. (2000) reported that ozone is not effective to disinfect wastewater down to 2 CFU/100 mL of total coliforms as attained by UV and PAA treatment to fulfil California standard although it meets WHO standard with 1000 CFU/100 mL of faecal coliforms. USEPA (2010) suggested that ozone Ct (concentration x time) of 1.8 (mg/L)-min is required to achieve 4-log reduction of viruses at low temperature (< 1 °C). Sigmon et al. (2015) later confirmed that the inactivation of *E. coli*, Coxsackievirus B5, Poliovirus 2, Adenovirus 2, φX174, and PRD-1 more than 4-log should be conducted using ozone Ct between 0.1 and 1.0 (mg/L)-min. Similar to the use of acids, ozone also has to be handled by experts given that it has strong oxidizing strength, making it less suitable for implementation in a decentralised treatment system where it is expected to be manned by non-experts.

Ultrafiltration disinfects the treated wastewater from a conventional sewage treatment system, achieving *E. coli* count to be lower than 10 CFU/100 mL for unrestricted water reuse purposes

1
2
3 according to Italy and US standards (Falsanisi et al., 2010). Mazuki et al. (2020) conducted a techno-
4
5
6 economic evaluation on UV, chlorination, microfiltration and ultrafiltration for water reuse. Chlorination
7
8
9 and UV integrated with Enviromulti-media (a biofilm media) were rated to be the most cost-effective
10
11
12 for restricted reuse purposes, while ultrafiltration is most economical for unrestricted reuse purposes.
13
14
15
16
17

18 Conventional technologies for water and wastewater treatment do not involve virion removal. The
19
20
21 reuse of treated water or wastewater resulted in an increasing concern on the high influential viral
22
23
24 loads during a pandemic. In a study by Wölfel et al. (2020), about 67% of faeces samples which
25
26
27 tested positive showed SARS-CoV-2 RNA counts near to the maximum counts sputum. More studies
28
29
30 detected SARS-CoV-2 RNA in wastewater samples (La Rosa et al., 2020; Randazzo et al., 2020).
31
32
33 The dose of chemical disinfectant required further adjustment during the pandemic. Since the
34
35
36 efficiency of tertiary treatment such as ozone, UV radiation, advanced oxidation remains unknown,
37
38
39 ultrafiltration is recommended as the porous membrane with pore size within 2 - 50 nm is sufficient to
40
41
42 remove SARS-CoV-2 with a diameter of ~100 nm (Bogler et al., 2020). This further justifies the
43
44
45 importance of having proper reclaimed water quality standards for decentralised treatment systems to
46
47
48 prevent sources of reclaimed water from becoming a threat for human consumption.
49
50
51

52 Out of all the different technologies outlined above, membrane technology seems to have the greatest
53
54
55 potential to be developed as a solution for decentralised water/wastewater treatment. This is due to
56
57
58 the versatility in customising the materials used to fabricate the membrane to suit the purpose of
59
60

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

reclaimed water - additional materials can be introduced to enhance certain functions (i.e. disinfection) of the treatment process. However, further investigations need to be done before any of these materials and technologies can be scaled up for practices of rainwater harvesting or greywater treatment and reuse.

4. IMPLEMENTED SOLUTIONS IN THE MARKET

4.1 Rainwater harvesting

Rainwater harvesting collects rainwater for reuse purposes onsite within the rainfall and hence stops contributing to surface runoff and minimises the risk of flooding in sewer systems. Rainwater harvesting systems range from using simple pails at the back of home, rain barrels (butts) to more elaborate designs with tanks, purification systems and pumps. Rainwater harvesting is one of the most cost-effective approaches to tackle the water scarcity obstacle. Different rainwater harvesting technologies are available in the market with the registered market value of around USD 9000 million in 2020 (Expert Market Research, 2020). According to the Global Rainwater Harvesting Market Outlook Report, the market value is projected to increase with a compound annual growth rate of 4.5% from 2021 to 2026 through analysis of global markets comprising North America, Europe, Asia Pacific and the rest of the world (Expert Market Research, 2020). Adoption of different rainwater

1
2
3 harvesting technologies across the global market for water reuse is based on 4 basic components
4
5
6 including catchment area, storage tank, treatment and recycling application.
7
8
9

10 11 12 13 14 *4.1.1 Catchment area* 15 16 17

18 Rainwater can be commonly harvested from rooftops and impervious surfaces. For the rooftop, the
19
20 rainwater follows the slanted roof to be collected at the end of the roof by a gutter that channels the
21
22 water through a vent pipe into storage tank (Pala et al., 2021). On the other hand, impervious land
23
24 runoffs include collecting rainwater from pavement, road and footpath areas. The catchment area
25
26 affects the quality of harvested rainwater. Water collected from roof areas will be of better quality than
27
28 others due to minimal contact with contaminants on land and thus, require less complex purification
29
30 processes (Samuel et al., 2012). Therefore, roof harvesting systems are commonly explored for
31
32 housing and industrial areas for non-potable uses, whereas surface runoff based harvesting systems
33
34 such as pond systems are widely utilised in farming land for irrigation purposes, where the water
35
36 quality requirement is less stringent (Zabidi et al., 2020). For the roof harvesting system, the material
37
38 used for roof and effective roof area will affect the quality of water and efficiency of water collection
39
40 respectively (Li et al., 2010). For example, Mostafa & Shafiuzzaman (2008) reported rainwater
41
42 collected from concrete roofs had higher suspended and dissolved solids, while galvanised roofs had
43
44 higher heavy metal content such as iron and zinc. Certain systems in the market specify the type of
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

roof material needed to be compatible with their systems especially for treatment facilities (Innovative Water Solutions [IWS], 2020).

However, in many low income regions including the remote locations and informal settlements, proper roofing for rainwater harvesting may be unavailable. The roof types in these regions range from mud roof, grass hut, thatched roof, pitched zinc roof with no interconnection and recycled-material composed roof. In these areas, different strategies have been adopted previously. In South Africa, mud dams, dug ponds, stone terracing and bed gardening are still used in many rural areas for water supply (Kahinda & Taigbenu, 2011). Nowadays, the simple concept of using a whole tank with gutter system as collection surface and storage container with no need for additional expenditure on piping equipment for channelling the rainwater from different catchment area makes the system easy for adoption in areas with no built-in water distribution piping systems (Heritage Tanks, 2016). Besides, for a single household with limited space available and low water requirement, a low cost inverted umbrella can be a feasible option (Hammed et al., 2017; Pala et al., 2021). In a study in Nigeria, rainwater collected from the umbrella harvester had lesser TSS (0.82 mg/L), TDS (32.97 mg/L), total hardness (10.67 mg/L) and nitrate (1.58 mg/L) than rainwater from galvanised roofs (TSS: 4.02 mg/L; TDS: 41.67 mg/L; total hardness: 36.33 mg/L; nitrate: 5.38 mg/L) (Hammed et al., 2017).

There is also an uptrend in man-made nature based catchments built specifically for rainwater harvesting. In Singapore, a similar umbrella concept but with advanced technology was introduced, that is Supertrees towers. The towers are composed of solar photovoltaic systems to capture solar energy, act as rainwater harvesters and function as a garden supporting growth of around 162, 900

plants (Solaripedia, 2014). Other catchments include green roof and green wall. In comparison to the general roof which acts as merely an impermeable capture area, a green roof with vegetation and substrate on it functions as catchment, storage and filtration system concurrently. Green roofs reduce the rate and volume of stormwater runoff. Green wall works on similar principles and is commonly adopted in commercial areas, improving the visual amenity of the building. Retrofitting of green walls is easier on commercial buildings due to accessibility of elaborate pipe networks for installation. In short, the choice of rainwater catchment area and its type are dependent on the fitting area and the installation place as a single-house (detached or semi-detached) residential area, apartments/ block of flats (housing estate) or commercial building. For any of these catchment states, an appropriate decentralised rainwater harvesting scheme can be found in the market but usually needs to be customised to obtain a cost-effective design and optimal performance of rainwater collection.

4.1.2 Storage tank

Storage tanks are constructed using different materials and come in various designs in the market. Metal, polyethylene, cement-brick and ferro-cement are generally used materials (Islam et al, 2010; Li et al., 2010). The size of the storage tank is chosen based on the size of the catchment area, level of rainfall and daily water requirement (Li et al., 2010). In general, storage tanks can be installed above ground and underground. Few advantages of above ground systems are lesser energy required to distribute water for reuse via gravity flow, easy detection of leakages and cleaning work (Zabidi et al.,

1
2
3 2020). However, the tank will occupy precious space around the house and not aesthetically pleasant.
4
5
6 For aesthetically appealing residential rainwater harvesting systems, there are a wide variety of
7
8
9 choices for designer rainwater vessels (modern, rustic, free-standing rainwater tank, wall-mounted
10
11
12 tank with or without integrated planter pot) in the market (Bacfree, 2021). On the other hand, the
13
14
15 underground rainwater harvesting system can save space around the house. However, it involves
16
17
18 higher cost due to underground digging and construction and more energy for distribution. To
19
20
21 overcome this disadvantage, there are tanks available with shallow dig design which mean less
22
23
24 ground required to excavate, less cost, easy maintenance and up to 80% less digging work (Graf,
25
26
27 2021; Stormsaver, 2021). While underground installation while might protect the tank from sunlight
28
29
30 induced evaporation, this will limit easy inspection for leakage especially in underground with a web of
31
32
33 tree roots that can damage the tank, has higher risk of contamination and need for higher power
34
35
36 pump to extract water for distribution (Zabidi et al., 2020). Storage tanks for different scales of
37
38
39 decentralised rainwater harvesting schemes are common in the market but new forms that are more
40
41
42 efficient and effective for installation and water quality protection fit for specific applications need to be
43
44
45 improved with further investigations in the future.
46
47
48
49
50
51

52 *4.1.3 Treatment and recycling*

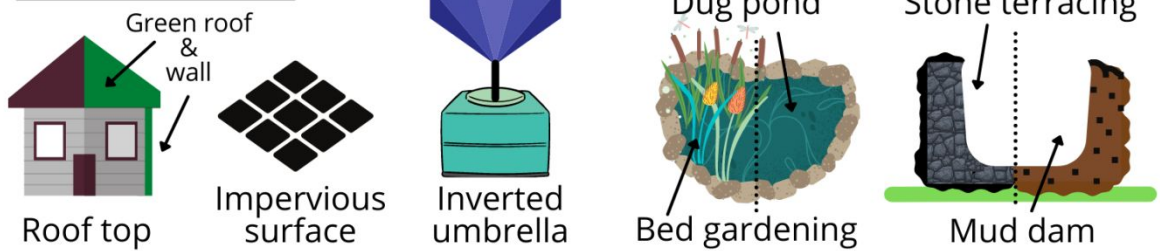
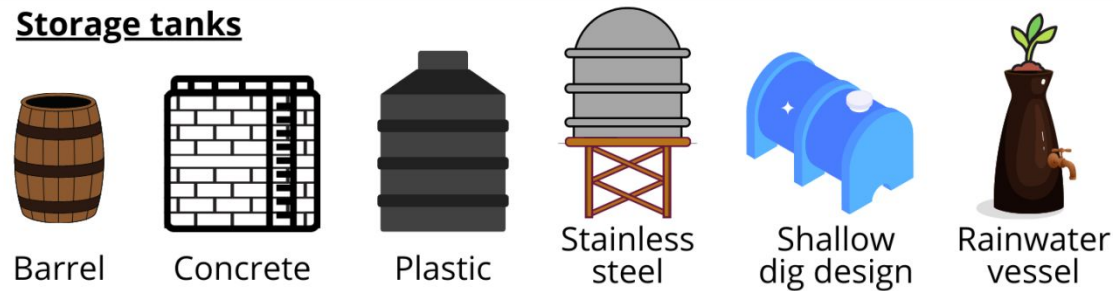
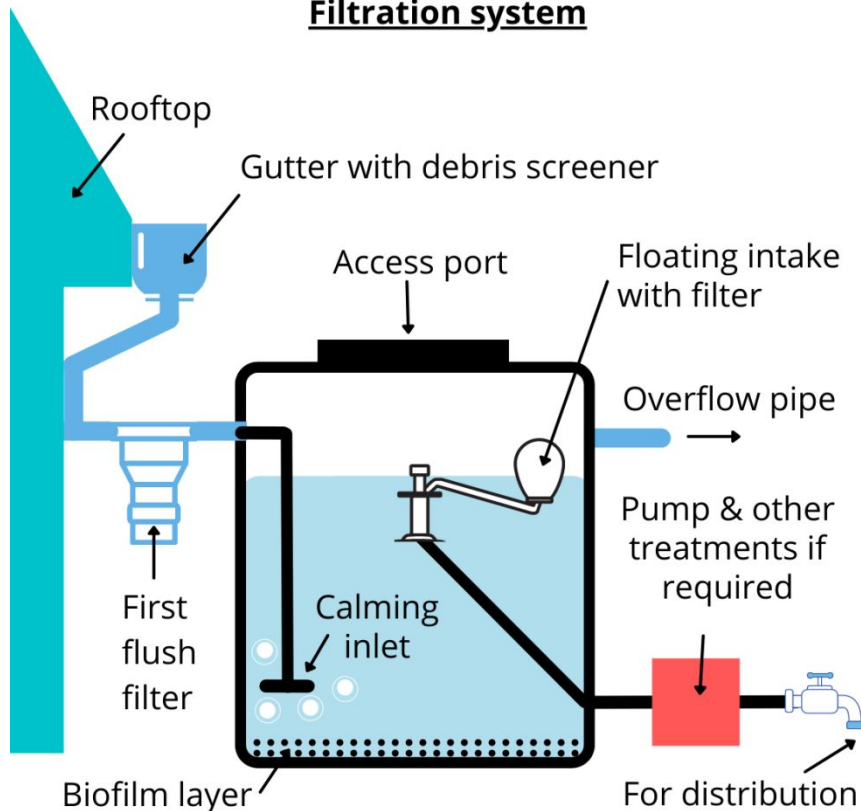
53
54
55 *Filtration system*
56
57
58
59
60

The rainwater from the rooftop might be contaminated with dust, dry leaves, bird droppings, microbes or even heavy metals corrodes from the roof material and has to be purified to varying degrees depending on its water reuse application. The filters used in the market range from a simple screener to sophisticated filters which will dictate the purpose of water reuse. Primary filters such as the first flush diverter or collector present in most rainwater harvesting systems as shown in Figure 3 (Heritage Tanks, 2016; IWS, 2020; WISY, 2020). It functions to discard the first collection of rainwater as there might be accumulated contaminants from dirt on the roof. There are two types of filters in the market for that function based on a unique concept. The vertical filtration system functions as both first flush and filter with 90% efficiency in collecting rainwater (WISY, 2020). The vertical filter mesh works based on the physics law of water, namely, adhesion, cohesion and gravity. The remaining 10% of the rainwater used to self-clean the device, thus preventing any blockage of the filter. The second filter is the dual intensity self-cleaning rooftop rainwater harvesting filter (D&D Ecotech Services, 2021; Water Field Technology, 2021). This self-cleaning device has efficiency above 90% to channel treated water into storage tanks (Rainwater Harvesting, 2015). The simple design of the filter allows its mounting to a wall and occupies very little space and can be used for both individual house and apartment-type households. There is no special requirement for the storage tank and thus allow for flexibility of using tanks available locally. In the storage tank, the predominantly adopted primary filtration mechanism is physical method via sedimentation and together with calming inlet to prevent the disturbance to the layer of healthy biofilm layer that formed by small quantities of sediment present in the rainwater despite the prior filtration process, overflow siphon to remove floating

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

particles such as pollen or dust forming a layer on top of the rainwater when the tank overflows and floating suction filtration in some products to ensure the intake of cleanest water from right below the water surface (Watts, 2017; IWS, 2020; WISY, 2020). The water can be distributed for garden irrigation or pumped for storage to be used later for non-potable water reuse such as toilet flushing, washing vehicles. In certain localities, activated carbon and charcoal filters which work on the principle of adsorption are easy and inexpensive to find are used to remove impurities such as heavy metals and reduce odour from rainwater, especially in low income regions (Islam et al., 2010).

For Peer Review

Catchment areas**Storage tanks****Filtration system****Other treatments**

UV disinfection

Photocatalytic

Chlorination

Membrane filtration

Ozonation

Slow sand filtration

Recycling applications

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

Figure 3. A summary of rainwater harvesting with four basic components including catchment areas, storage tanks, treatments and recycling applications.

Natural filtration

Apart from the filtration systems discussed above, nature-based filtrations are also gaining popularity such as green roofs and walls. Green roofs can be either extensive or intensive for functional purposes only or for aesthetic look too respectively (Laminack, 2014). Extensive green roofs are almost five times shallower than intensive roofs and covered with stress tolerant and smaller plants (require less growth media) (Laminack, 2014). Besides functioning as a rainwater sponge and reducing the burden of stormwater drainage, the green roofs can improve a few quality parameters of rainwater such as reducing its acidity and nitrate-ammonium content (Laminack, 2014). However, green infrastructure harvesting methods in general produce rainwater of lesser quality than those collected by rooftops. Green wall works on similar principle but follows vertical structural design. The collected rainwater can be used directly for irrigation of commercial and residential buildings' landscapes, besides toilet flushing.

Secondary treatments

In order to convert harvested rainwater to potable water, sophisticated treatment is required such as UV disinfection, photocatalytic, chemical treatment (e.g. chlorination), membrane-based filtration, ozonation and slow sand filtration (Buntat et al., 2015; Omar et al., 2017; Lani et al., 2018). Suitability

of the treatment method depends on the water quality requirement for reuse and cost of the treatment system. In general, the cost of operation and maintenance of a rainwater harvesting system is economical if a simple water purification system is chosen for non-potable uses. Successful implementation of the rainwater harvesting system is multifactorial including tax relaxation, local authority incentives and public awareness. For example, in Uganda subsidies are provided for rural households to buy rainwater harvesting system construction items, while in Germany, stormwater taxes are exempted for premises that practise rainwater harvesting (Lani et al., 2018). While rainwater harvesting is gaining its popularity, greywater recycling potential for water reuse is also tapped worldwide.

4.2 Greywater recycling

Compared to rainwater harvesting that is dependent on the weather pattern and specifically wet periods, greywater is a more reliable alternative as a continuous water supply source. Greywater recycling systems range from very basic to advanced methods of recycling. Unlike rainwater harvesting that can be used without any treatment, the collected greywater needs to undergo a certain level of treatment before it can be used even for non-potable purposes. Some of the commonly used treatment techniques include physicochemical (filtration, adsorption, reverse osmosis) and biological (membrane biological reactor, rotating biological contactor, biological aerated

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

filter, waste stabilisation ponds) methods (Sivakumar & Ramezaniapour, 2012). The greywater recycling market is driven by increased need to look for freshwater resources alternatives, government incentives and public awareness. The market is predominantly segmented based on the treatment systems and applications. Direct reuse of collected greywater is seldom used in commercial settings and the concept of laundry-to-landscape is more prevalent in residential areas. Unlike rainwater, collected greywater has to be reused promptly since chemical parameters such as BOD and COD can change significantly within a short time period (Li et al., 2010). The British Standard BS8525-1:2010 outlining the Greywater Systems Code of Practice advised against using greywater for applications requiring higher water quality including bathing and dishwashing (BSI, 2010). The technologies used for greywater treatment can be classified under three systems (one-stage, two-stage and multi-stage) that are critically analysed below.

4.2.1 One-stage system

To date, there are a few companies producing one-stage process system, where the water is reused within a short residence time to prevent significant changes to the chemical parameters of greywater and only minimal treatment necessary (Aqua2use, 2018a; Aquarius Wastewater Management Systems [AWMS], 2020a, 2020b). Coarse filtration is the common type of one-stage process system – typically using in-line bag filters, mesh filters and filtration media (Pure Water LLC, 2010; Matala Water Technology Co., Ltd, 2011; Hitchner, 2016). The filter pore size ranges from about 100 µm to

even as low as 5 μm and dictates the quality of treated water produced. When influent passes through the filtration unit, the suspended solids are trapped building the biofilm. It comes in various surface areas and shapes. Increasing the surface area of the filter will improve its sludge holding capacity. One-stage treatment system is a compact above ground system and can be powered by normal domestic electrical supply (Aqua2use, 2018a; AWMS, 2020b). It is available at small scale such as small homes and the main application of water reuse is almost always for lawn and garden irrigation only. The only maintenance work required is periodic cleaning of the filters. Green technologies are also used in greywater recycling. One example is the constructed wetland as a low cost and low impact system, especially for rural areas or small communities in developing countries (Oladoja, 2017; Collivignarelli et al., 2020). Vertical flow constructed wetland functions as one continuous physical and biological filtration system without the need of a preliminary settling tank (Oladoja, 2017; Goncalves et al., 2021). Native species of wetland plants can be chosen for water purification. While the household greywater reuse is more prevalent among the developed countries, this concept is still at its elementary stage and on-site decentralised systems such as wetlands is recognised as appropriate technology in under-resourced countries (Oladoja, 2017). Constructed wetlands such as reed beds are also adopted in developed countries due to dual functions of the eco-pond which are captivating green landscape and sustainable water recycling systems (Stefanakis, 2020). Other examples of green technologies are using natural filtration systems such as green roofs and walls which serve the dual purpose of urban cooling and water recycling.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

4.2.2 Two-stage system

Filtration and disinfection

Two-stage system consisting of filtration and disinfection is the most common technology used for domestic greywater reuse. There are products in the market that utilise coarse filtration and chemical disinfection, coarse filtration and non-chemical disinfection, membrane filtration and chemical disinfection and multi-media filtration and chemical disinfection (Pure Water LLC, 2010; Hitchner, 2016; AWMS, 2020c; Aquaco Water Recycling Ltd [AWR], 2021a, 2021b; Greyter, Inc, 2021a, 2021b). Coarse filtration normally involves metal strainer, mesh/ gauze filter, bag filters, organic clay filter, polymeric filtration media, and disc filter. Ultrafiltration membrane technology produces better quality treated greywater due to smaller pores with size ranges from 0.01 μm to 0.1 μm . However, higher pressure is needed to function hence higher energy consumption is required compared to other conventional filtration technology (Li et al., 2010). Another factor restricting its economic viability is membrane fouling that reduces the efficiency of membrane filtration (Ezugbe & Rathilal, 2020). Thus, greywater treatment systems in the market commonly use hybrid membrane processes where other purification methods are used as pre-treatment to reduce membrane fouling. In contrast, multi-media filters consist of different layers of media such as sand, coal and gravel which are of varying size to purify water. The multi-media greywater system is often targeted at the Middle Eastern market that conventionally uses sand filtration treatment methods for rapid filtration (AWR, 2021a). This system has a very small footprint, fast filter rate, low maintenance and running cost (AWR, 2021a). Chemical disinfection is achieved by dosing of chlorine, bromine or sodium hypochlorite solution

(Pure Water LLC, 2010; AWR, 2021a, 2021b; Greyter, Inc, 2021a, 2021b). Other chemical free disinfection processes in the market are UV disinfection, ozonation and electrolysis (Hitchner, 2016; AWMS, 2020c; Aqua2use, 2018b). The applications of greywater treated by filtration and disinfection processes are mostly limited to lawn irrigation and toilet flushing.

Filtration and biological treatment

For applications that require a more stringent water quality such as laundry, filtration alone or filtration and disinfection may not produce treated water that meets the quality standards. Hence, biological treatment is required in these cases to complement the filtration process to remove the organic matter in the greywater. There are few treatment systems in the market that combine mechanical and biological treatments such as membrane bioreactor (MBR) (Aqua2use, 2018b; Aquacell, 2021; AWR, 2021c). MBR is a combination of an activated sludge reactor and a filtration unit. Activated sludge produced via aeration to create ideal conditions for breakdown of biodegradable material. A line of filtration units available in the market range from polymeric biofilter, ultrafiltration membrane and microfiltration membrane (Aqua2use, 2018b; Aquacell, 2021). Even though this system has higher energy demand compared to basic filtration-disinfection systems, it produces treated water with higher quality that can be used additionally for washing clothes. Routine membrane cleaning is required for the system to operate optimally. MBR are available at different scales and can be configured into available space just like other technologies.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

4.2.3 Multi-stage system

Technologies in the market could not be easily categorised as one-stage or two-stage systems because a complete system almost always comprises multiple treatment process units, especially those to be used for higher quality water production. Sedimentation tanks are present in almost all greywater treatment systems and precede any other processes. Other physicochemical methods used are activated carbon, coagulation, aeration, and clarification. In general, the cost of the greywater recycling system is dependent on sophistication of the technology, water reuse application, maintenance and ease of retrofitting/ renovation needed (Li et al., 2010). Right technology for a particular chosen application will ensure substantial financial savings. Advanced treatment technology for just basic use such as lawn irrigation in a residential area might not be economically viable.

There are ever growing technological advancements of water-saving technologies in the market. However, there are technologies that failed to scale up despite widespread promotion due to the lack of flexible water management framework, changing dynamics of the competitive water market, pricing, and social constraints to adopt those technologies, especially this is the case in the urban settings. This requires that a comprehensive conceptual framework is developed for urban water reuse that will be further discussed in the next section.

5. CONCEPTUAL FRAMEWORK OF WATER REUSE STRATEGIES

Understanding of current technological developments of water reuse in the market is necessary for the technology adoption and incorporating into urban water systems. However, having access to these technologies per se cannot guarantee their adoption and applications by the public or water utilities. This is mainly due to the fact that the main dimensions of the sustainability framework (i.e. economic, social, environmental, governance and asset aspects) need to be satisfied when introducing strategies based on these technologies (Behzadian & Kapelan 2015a). Hence, development of urban water reuse strategies first requires a conceptual framework to include all relevant metrics or key performance indicators (KPIs) derived from the assessment criteria taken from the sustainability framework. Note that creating an exhaustive list of KPIs for water reuse technologies can be challenging and hence the decision makers should select only a limited number of those KPIs that are of interest to the stakeholders and the public. Some examples of these KPIs for water reuse schemes in urban water systems are capital or operation and maintenance costs, efficiency, total contaminations, reliability/resilience/vulnerability of water supply, flood volume, greenhouse gas emissions, public acceptance, usability, etc. These KPIs can be quantified by either quantitative methods such as water balance conceptually- or physically-based models or qualitative methods such as surveys or questionnaires filled out by the public, experts or stakeholders. Various techniques have been used in the literature for assessment of water reuse-based technologies based on a number of defined KPIs such as reliability assessment of industrial wastewater reuse by using combined event tree and fuzzy fault tree analysis (Piadeh et al. 2018). After the assessment of potential water reuse strategies for all KPIs individually, the strategies need to be analysed and compared through either

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

ranking the strategies with respect to all analysing KPIs (Morley et al. 2016) or a scenario-based analysis to track all plausible consequences of the analysing strategies (Momeni et al. 2021). This can be performed via a suitable multi-criteria decision analysis (MCDA) such as CP (compromise programming) method (Behzadian and Kapelan 2015a) or fuzzy AHP (analytical hierarchy process) method (Ardeshir et al. 2014).

Figure 4 shows a sample of a conceptual framework for water reuse decision making. This framework can be based on three main modules including 1) problem definition; 2) impact assessment and 3) MCDA. In the first module, water reuse technologies are selected and hence relevant strategies are formed. Evaluation criteria derived from the sustainability framework are also determined mainly based on the stakeholder preferences. Within the second module, KPIs are identified and the water reuse strategies are evaluated by either quantitatively for quantitative KPIs or qualitatively for qualitative KPIs. The impact assessment will feed the population matrix in the third module in which the strategies are either ranked by using an MCDA method or analysed based on the plausible scenarios.

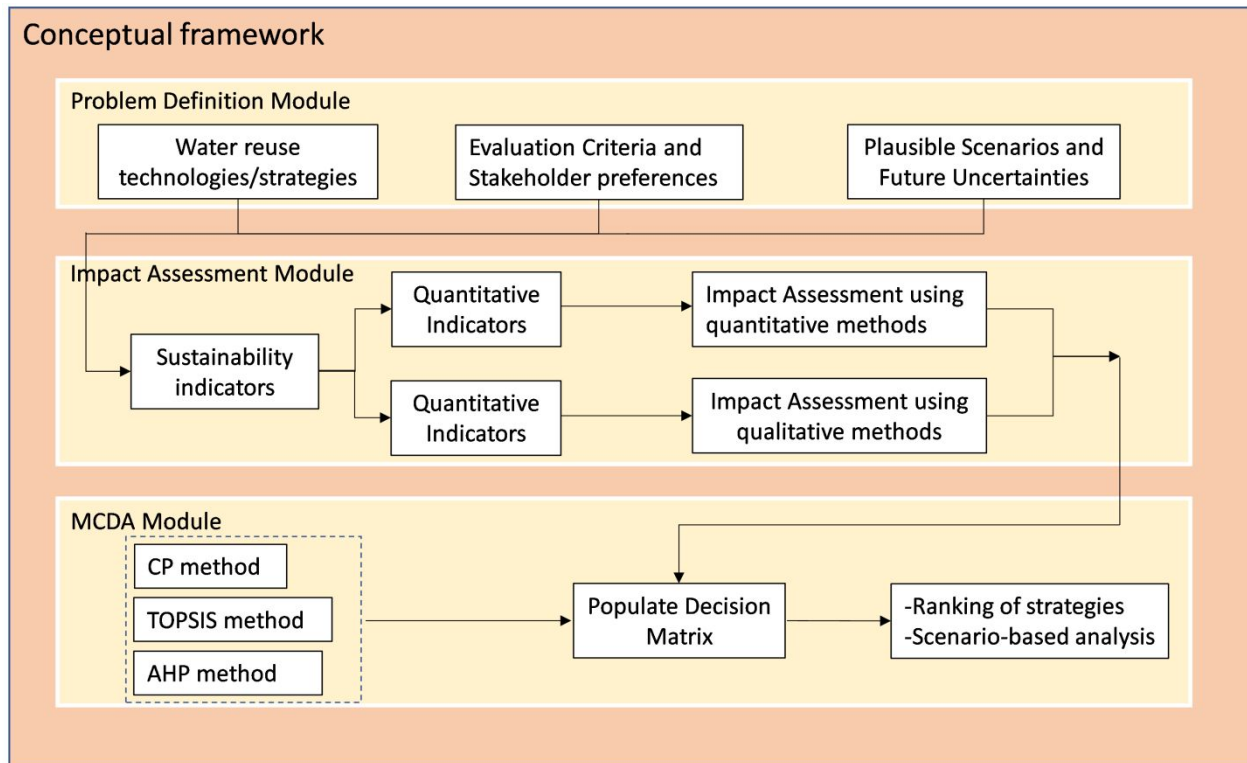


Figure 4. Conceptual framework suggested for water reuse decision making.

5.1 Why is urban water management important?

In addition, urban areas, the hotspots of water use worldwide, are facing water sustainability obstacles including fragmented urban water management. Water security is defined as “the capacity of a population to safeguard sustainable access to adequate quantities of acceptable quality water for sustaining livelihoods, human well-being, and socio-economic development, for ensuring protection against water-borne pollution and water-related disasters, and for preserving ecosystems in a climate of peace and political stability” (UN Water, 2013). The Sustainable Development Goals (SDG) were adopted by all the United Nations Member States in 2015 with 17 SDGs, which calls for global action

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

in reducing inequalities to end poverty and other deprivations (UN, 2020). In particular, SDG number 6 is to ensure availability and sustainable management of water and sanitation for all.

Despite the progress since 2000, 2 billion people (26% of the global population) still have no access to safely managed drinking water services and 29% of the world population have no basic hand washing facilities with soap and water in their dwellings (UN-Water, 2021). While health is a central quality of life indicator, poor health is often contributed by inadequate provisioning of basic needs such as safe, affordable and accessible water, especially in developing countries. Unsafe sanitation contributed to 775,000 deaths annually (Ritchie & Roser, 2021b). In 2020, 46% of the world population had no access to safely managed sanitation and 6% of the world was practising open defecation (Ritchie & Roser, 2021b). Childhood diarrhoea is significantly correlated with poor sanitation, poor hygiene or unsafe drinking water (UNICEF & WHO, 2019). Consumption of untreated and contaminated water due to unsanitary waste disposal has resulted in 1.7 billion diarrhoeal diseases in children under 5 years old and in 2017, 533,768 children died of diarrhoeal diseases (WHO, 2017; Dadoinate et al., 2019). Accessibility to safely managed water is a wicked problem. There is no direct strategy to solve this problem as there are few interlinked challenges acting as barriers, including growing water demand and scarcity, water pollution and the impact of climate change on water availability.

Population expansion, climate change and urbanisation are putting strain on urban water infrastructure systems of cities worldwide, especially in developing countries. According to UNESCO

(2018), there are 763 million internal migrants which is more than three times that of 258 million foreign migrants, globally. The disparity in a monthly income of the urban and rural populations has led to the megatrend of internal migration of the rural communities to the metropolitan cities. For example, in Indonesia, the world's fourth most populous nation, almost 50% of the population reside in urban areas, where rural-urban migration is top of the charts (UNESCO, 2018). Urban migration goes along with the vast expansion of informal settlements, congestion, growing food demand, water resources degradation and inappropriate management of water, sanitation and waste disposal services (International Water Association, 2018). These challenges emphasise the need for sustainable urban water development now more than any previous time.

The goal of urban water development is to create adaptable, comfortable living, productive, and sustainable townships. One important element to address water-related concerns is the environment-water interface, thus nature-based solutions might provide a more sustainable water use efficiency (UNESCO, 2021). To close the water loop of cities for better water sustainability, diversifying the water sources through maximising water collection by capturing every drop of rain that falls in the land and rendering water an endlessly renewable resource through wastewater reclamations are among the strategies that could be adopted. The technologies introduced in this study provide opportunities to form intervention strategies based on harvesting rainwater and reusing greywater for future planning of urban water systems. This is important because the assessment of intervention strategies for integrated urban water systems shows those strategies that include water reuse technologies are ranked high and more advantageous than other conventional ones (Behzadian & Kapelan, 2015a,

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

2015b). Furthermore, as the demand for clean, safe water increases, these technologies that are potentially applicable for decentralised schemes can offset the need for solely relying on centralised water supply. An understanding of the benefits of each technology and the resultant output for use is therefore crucial in designing a flexible system for reuse, capable of adapting to various local and climatic conditions.

Table 4 shows a number of sustainability indicators for two types of water reuse technologies with their options. This synthesis was arrived at by aggregating and mapping rainwater and greywater technologies and its application across the major sources (household- individual and urban- local) to mainly show the potential of these technologies and the ability to expand these to similar infrastructure types. The hotspots for the usability of the technological tools in urban setting can be categorised into residential, commercial and industrial areas, urban farming and informal settlements which will be briefly discussed in the next section.

Table 4. A number of sustainability indicators for rainwater and greywater technologies with their supply expansion potential (Information on filtration, technology and source adopted from Boano et al., 2020). Note that water expansion potential refers to the adaptability and ease of development in communities, leading to the potential of a higher rate of use and the ability to expand or increase the water supply in developing communities.

Type of water reuse and treatment	Specific technology	Scale		Source of water reuse	Usability/ Place of consumption	Water supply expansion potential
		Household	Urban			
RAINWATER						
Nature based filtration	Constructed Wetlands/bioswales	X	X	Street runoff, Roof	Agriculture, Recreation, Toilets flushing	High
	Green wall	X		Roof	Toilets flushing/agriculture	Low
	Green roof	X		Roof	Toilets flushing/agriculture	Medium
Physical filtration	Sedimentation tank + sand filter	X		Street runoff, Roof	Toilets flushing/agriculture	Medium
	Sedimentation tank + granular active carbon	X		Street runoff, Roof	Toilets flushing/agriculture	Low
GREYWATER						
Nature based filtration	Constructed Wetlands/bioswales	X	X	Shower, sinks, bath, kitchen	Toilets flushing/agriculture	High
	Green wall	X	X	Shower, sinks, bath, kitchen	Toilets flushing/agriculture	Low
	Green roof	X	X	Rainwater	Toilets flushing/agriculture	Medium
Physical filtration	Activated charcoal (charcoal)	X	X	Bath, laundry	Toilets flushing/agriculture	Medium
	Sedimentation tank + sand filter	X	X	Kitchen	Toilets flushing/agriculture	Medium
	Hybrid granular active carbon electrochemical system	X	X	Shower/bath	Toilets flushing/agriculture	Medium
	Sedimentation tank + granular active carbon	X	X	Kitchen	Toilets flushing/agriculture	Medium
Chemical filtration	Coagulation		X	Shower, sink/bath	Agriculture	High
	Electrocoagulation		X	Shower, sink, kitchen	Agriculture	High
	EC/O3/UV		X	Shower, sink	Agriculture	High
	Photocatalytic fuel cell		X	Laundry	Agriculture	High
Bio filtration	Rotating biological contactor		X	Laundry, bath, kitchen	Agriculture	High

Moving bed biofilm reactor	X	Laundry, bath	Agriculture	High
Membrane bioreactor	X	Laundry, bath, kitchen	Agriculture	High
Up-flow anaerobic sludge blanket	X	Shower	Agriculture	High
Biological activated membrane bioreactor	X	Bath, sink	Agriculture	High

Abbreviations: EC- electrocoagulation, O3- ozonation, UV- ultraviolet.

For Peer Review

5.2 The hotspots for water reuse in urban setting

The traditional urban water management consists of public water system connecting the centralised water treatment plant to end-users, sewerage system (in certain areas) and finally, wastewater treatment plant (if available) to discharge treated or untreated wastewater as shown in Figure 5. The design of the traditional urban water management follows linear economy, where freshwater resources are taken, treated into potable water quality, used and released back into the environment as low quality wastewater. In order to accelerate the progress towards achieving sustainable urban water development globally, circular economy should be adopted for a more holistic water resources management. One way to achieve this is by practising water reuse via rainwater harvesting and greywater recycling in a small or large scale (Figure 5). The circular economy concept can be applied to a whole country or to particular hotspots in urban areas.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

For Peer Review

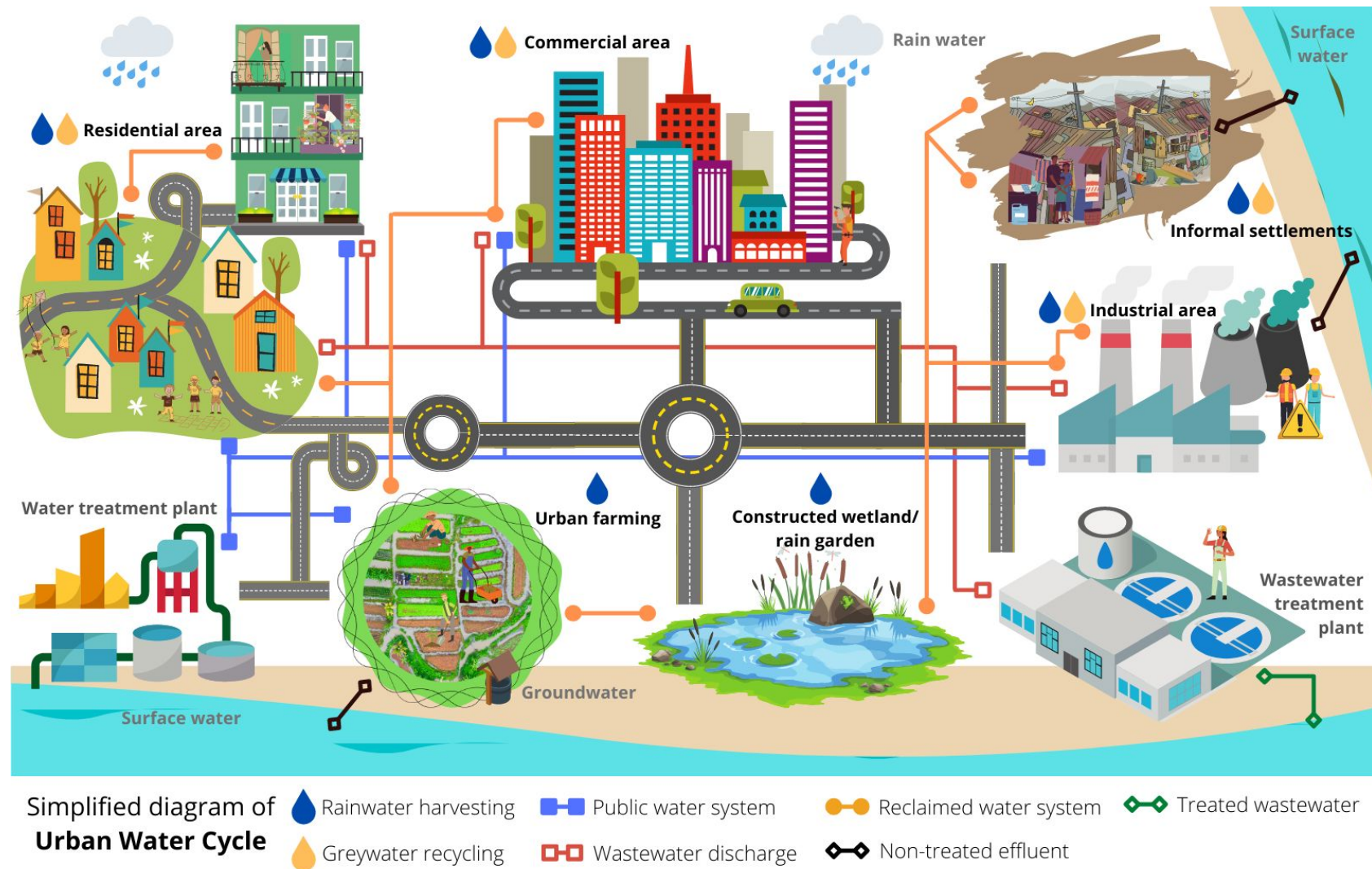


Figure 5. A simplified diagram of urban water cycle, where rainwater harvesting and greywater recycling can be adopted in addition to conventional centralised water supply in urban water management

5.2.1 Residential, commercial and industrial areas

Residential, commercial and industrial areas are inherently part of the urban water management and supplied with the municipal water systems. The public water system is commonly sourced from depleting freshwater sources. Decentralised systems would help to reduce the dependence on the centralised water supply. Residential areas could benefit from both rainwater harvesting and greywater recycling. It will be advantageous to mix the two water sources because sometimes the amount of rainwater is insufficient to provide supply for non-potable reuse and greywater is consistently available. The treatment of rainwater together with greywater will lower the concentration of materials that needs to be removed. But the inconsistent supply of rain may imply that the system has to be robust enough to deal with occasional changes in feed characteristics. Reclaimed water in these localities has low to medium water supply expansion potential and thus more commonly utilised for on-site toilet flushing and garden irrigation. Nature-based and physical filtration technological tools are more suitable to produce water for reuse. Capitalising on water-saving strategies which are focused on nature-based filtration can therefore serve as a ubiquitous and accessible method for many socio-economically disadvantaged communities to expand their water supply. Constructed wetlands and bioswales are also gaining popularity in the latest townships. Green walls and roofs also can be retrofitted into larger buildings, i.e.: condominium and commercial buildings. Besides helping to save water, these green technologies add aesthetic value. When coupled with water-saving fixtures and appliances, and efficient irrigation technologies and practises, these filtration technologies can increase clean water accessibility significantly with high community acceptance in these areas.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

5.2.2 Urban farming

The United Nations World Water Development Report (2021) stated one of the water management strategies for food production is obtaining water for irrigated agriculture, mainly from nature-based and non-conventional sources (UNESCO, 2021). In communities where reclaimed greywater is still not accepted for domestic use, the recycled water can be channelled to the nearest urban farms. Urban farming is gaining popularity in countries like India, Ghana, Kenya and Vietnam, to name a few. For many developing countries, agriculture contributes a smaller share of income compared to industrial sectors, it however, consumes largest volumes of water, contributing to 70% of freshwater withdrawal globally (Ritchie & Roser, 2017). Indonesia is the largest agricultural water consumer in Southeast Asia, at nearly 92.76 billion m³ , followed by Vietnam and Philippines (Ritchie & Roser, 2017). Urban and peri-urban farms produce 80% and 40% of the total of fresh vegetables and eggs respectively in Hanoi, Vietnam (Corbould, 2013). In Ghana's capital, almost 90% of consumed vegetables are produced within the city (Corbould, 2013). Improvement in crop productivity with a lower percentage of freshwater use should be the way forward and reclaimed water could be a sustainable solution. Recycled water used for vegetative farming has to meet certain quality standards, thus additional technological tools such as chemical filtration and bio filtration are needed. In general, reclaimed greywater for agricultural use has high water supply expansion potential and rainfed irrigation is also feasible in-house for areas with sufficient rainfall.

5.2.3 Informal settlements

In most developing countries, informal settlements are currently not part of a city's urban water management system. The lack of water supply facilities results in the informal settlers' reliance on self-fetched surface water, purchase of water from water kiosks and self-made unsustainable water systems. There are ongoing social challenges such as poor drinking water quality and incompetently built sanitation infrastructure. The greywater generated by them on-site would harbour germs and cause a hygiene-related disease burden to the congested-living communities. In areas with frequent rain, simple medium-sized rainwater tanks with gutter or umbrella harvester and filter systems could provide them with reasonably clean water for non-potable uses or even potable purposes with simple disinfection processes such as boiling or solar disinfection.

However, when it comes to decentralised greywater recycling, governance, the fragmented and fragile geography, lack of human resources (locally trained runners for maintenance services are lacking), in addition to the disadvantaged socio-economics of the local population obstruct the implementation of sophisticated technologies (Putri, 2017; Prescott et al., 2021). In this situation, the best practice is to adapt a modular onsite system, especially in areas with limited space, with locally available materials, and green infrastructure is a cost-effective option (Putri, 2017; Prescott et al., 2021). Phytoremediation technology using simple constructed wetlands aided with simple sediment and carbon filter are successful examples of decentralised systems adopted in many informal settlements such as Lima (Peru) and Machaki village (Pakistan) (Prescott et al., 2021). With a basic

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

water piping infrastructure facility, decentralised greywater recycling could be done to produce water for mobile toilets too.

Water literacy: Is it essential?

Literate environments are necessary to pursue water management sustainability, especially when it comes to the provision of potable water. Informal settlements are often correlated with poverty. Poverty is a wicked problem and tangled with other challenges such as lack of education and inequalities (Maniam et al., 2021). Moreover, the adaptation of water sustainability is effortful and time-consuming. The current roadmap for implementation of the SDG itself is planned over a 15-year period, while the Millennium Development Goals had been in place from 2000 to 2015. The same goes to behavioural change among the individuals to be more water literate, it is a long-winded commute. Any water supply and sanitation projects carried out in these localities or other areas should be followed by water education programmes. This is because any unhygienic practice at the point-of-consumption would make the effectiveness and efficiency of water recycling procedure insignificant.

Conclusion

Anthropogenic activities have led to water stress in many parts of the world. With limited freshwater sources, reclamation of non-conventional water sources such as rainwater harvesting or greywater treatment and reuse provide a viable alternative to alleviate water stress. There are various

technological options available to carry out rainwater harvesting or greywater treatment and reuse. A conceptual framework is proposed to provide an overview on the applicability of different technologies under various settings, mapped against usability of the treated water and the potential for water supply expansion. Decision makers could refer to the conceptual framework to decide on a suitable decentralised solution depending on limiting criteria. Decentralised systems of rainwater harvesting and greywater treatment for reuse are expected to play an important role to reduce the dependence on the centralised water supply, contributing to water supply for urban farming and even to informal settlements.

Funding Information

The authors would like to acknowledge the Royal Academy of Engineering, Frontiers of Engineering for the financial support (under the Frontiers of Engineering for Development scheme with contract no FoE2021\9\2) to conduct this work. The authors wish to acknowledge the constructive comments made by the reviewers that significantly improve the quality of the review study.

References

Abdel-Shafy, H., Mansour, M. S. M. & Al-Sulaiman, A. M. (2019). Anaerobic/aerobic integration via uasb/enhanced aeration for greywater treatment and unrestricted reuse. *Water Practice & Technology*, 14(4), 837-850. <https://doi.org/10.2166/wpt.2019.065>

- Abuelfutouh, N. A. K., Jamie, M., Nour, A., & Fuad, N. I. (2020). Rainwater harvesting quality assessment and evaluation: iium case study. *IUM Engineering Journal*, 21(1), 12–22. <https://doi.org/10.31436/iiumej.v21i1.1139>
- Abulude, F. O., Ndamitso, M. M. & Abdulkadir, A. (2018). Rainfall water quality assessment in atmospheric deposition of an urban area” a case study of Akure in Nigeria. *Anthropogenic Pollution Journal*, 2(2), 1-9. <https://doi.org/10.22034/ap.2018.562429.1019>
- Ahuja, S. (2019). Global Water Challenges and Solutions. In: S. Ahuja, ed. *Advances in Water Purification Techniques* (pp. 17-39). United States: Elsevier.
- Aiken, S. R. & Leigh, C. H. (2015). Dams and indigenous peoples in Malaysia: development, displacement and resettlement. *Geografiska Annaler: Series B*, 97(1), 69-93.
- Alcalde-Sanz, L. & Gawlik, B. M. (2017). JRC Science for Policy Report: Minimum quality requirements for water reuse in agricultural irrigation and aquifer recharge. Luxembourg: European Union.
- Al-Hamaiedeh, H. & Bino, M. (2010). Effect of treated grey water reuse in irrigation on soil and plants. *Desalination*, 256(1-3), 115-119. <https://doi.org/10.1016/j.desal.2010.02.004>
- Alrousan, D. M. A. & Dunlop, P. S. M. (2020). Evaluation of ozone-based oxidation and solar advanced oxidation treatment of greywater. *Journal of Environmental Chemical Engineering*, 8(104309), 1-8. <https://doi.org/10.1016/j.jece.2020.104309>
- Al-Wasify, R. S., Ali, M. N. & Hamed, S. R. (2018). Greywater treatment for irrigation purposes using pottery scraps and aerated moving bed biofilm reactor. *Desalination and Water Treatment*, 126, 40-47. <https://doi.org/10.5004/dwt.2018.22897>
- Angelakis, A. N., Asano, T., Bahri, A., Jimenez, B. E., & Tchobanoglous, G. (2018). Water reuse: from ancient to modern times and the future. *Frontiers in Environmental Science*, 6(26), 1-17. <https://doi.org/10.3389/fenvs.2018.00026>
- Antonelli, M., Mezzanotte, V. & Nurizzo, C. (2008). Wastewater disinfection by UV irradiation: Short and long-term efficiency. *Environmental Engineering Science*, 25(3), 363-373. <https://doi.org/10.1089/ees.2006.0268>

- Antonopoulou, G., Kirkou, A. & Stasinakis, A. S. (2013). Quantitative and qualitative greywater characterization in Greek households and investigation of their treatment using physicochemical methods. *Science of the Total Environment*, 454-455, 426-432.
<https://doi.org/10.1016/j.scitotenv.2013.03.045>
- Aqua2use. (2018a). GWDD. [Online] Available at: <https://www.aqua2use.com/gwdd-landscape-garden-green-in-drought/> [Accessed: 13 August 2021].
- Aqua2use. (2018b). GWTS1000. [Online] Available at: <https://www.aqua2use.com/gwts1000-green-building-grey-water-recycle-treatment-methods/> [Accessed: 13 August 2021].
- Aquacell. (2021). Greywater Systems. [Online] Available at: <https://aquacell.com.au/commercial-water-recycling-systems/greywater/> [Accessed: 13 August 2021].
- Aquaco Water Recycling Ltd. (2021a). Multi Media System. [Online] Available at: <https://www.aquaco.co.uk/grey-water-recycling/multi-media-system/> [Accessed: 13 August 2021].
- Aquaco Water Recycling Ltd. (2021b). Residential. [Online] Available at: <https://www.aquaco.co.uk/grey-water-recycling/residential/> [Accessed: 13 August 2021].
- Aquaco Water Recycling Ltd. (2021c). MBR System. [Online] Available at: <https://www.aquaco.co.uk/grey-water-recycling/mbr-system/> [Accessed: 13 August 2021].
- Aquarius Wastewater Management Systems. (2020a). GREYWATER DIVERSION SYSTEM. [Online] Available at: <https://www.aquariuswastewater.com.au/wp-content/uploads/2015/06/Greywater-Brochure.pdf> [Accessed: 13 August 2021].
- Aquarius Wastewater Management Systems. (2020b). AQUARIUS "GECKO": Single Fixture Greywater Diversion Unit. [Online] Available at: <https://www.aquariuswastewater.com.au/wp-content/uploads/2015/06/Gecko-Brochure.pdf> [Accessed: 13 August 2021].
- Aquarius Wastewater Management Systems. (2020c). COMMERCIAL GREYWATER TREATMENT SYSTEM. [Online] Available at: <https://www.aquariuswastewater.com.au/our-products/commercial/greywater-treatment-system/> [Accessed: 13 August 2021].

- Ardeshir, A., Mohseni, N., Behzadian, K. & Errington, M., (2014). Selection of a bridge construction site using fuzzy analytical hierarchy process in geographic information system. *Arabian Journal for Science and Engineering*, 39(6), 4405-4420. <https://doi.org/10.1007/s13369-014-1070-2>
- Bacfree. (2021). Rainwater Harvesting System. [Online] Available at: <https://bacfree.com.my/professional/rwhs/> [Accessed: 13 August 2021].
- Barzegar, G., Wu, J. & Ghanbari, F. (2019). Enhanced treatment of greywater using electrocoagulation/ozonation: Investigation of process parameters. *Process Safety and Environmental Protection*, 121, 125-132. <https://doi.org/10.1016/j.psep.2018.10.013>
- Behzadian, K. & Kapelan, Z. (2015a). Advantages of integrated and sustainability based assessment for metabolism based strategic planning of urban water systems. *Science of the Total Environment*, 527, 220-231. <https://doi.org/10.1016/j.scitotenv.2015.04.097>
- Behzadian, K. & Kapelan, Z. (2015b). Modelling metabolism based performance of an urban water system using WaterMet2. *Resources, Conservation and Recycling*, 99, 84-99. <https://doi.org/10.1016/j.resconrec.2015.03.015>
- Bodnar, I., Szabolcsik, A., Baranyai, E., Uveges, A., & Boros, N. (2014). Qualitative characterization of household greywater in the northern great plain region of Hungary. *Environmental Engineering and Management Journal*, 13(11), 2717-2724. <https://doi.org/10.30638/eemj.2014.302>
- Boano, F., Caruso, A., Costamagna, E., Ridolfi, L., Fiore, S., Demichelis, F., Galvão, A., Pissoneiro, J., Rizzo, A. & Masi, F. (2020). A review of nature-based solutions for greywater treatment: applications, hydraulic design, and environmental benefits. *Science of the Total Environment*, 711(134731), 1-26. <https://doi.org/10.1016/j.scitotenv.2019.134731>
- Bogler, A., Packman, A., Furman, A., Gross, A., Kushmaro, A., Ronen, A., Dagot, C., Hill, C., Vaizel-Ohayon, D., Morgenroth, E., Bertuzzo, E., Wells, G., Kiperwas, H. R., Horn, H., Negev, I., Zucker, I., Bar-Or, I., Moran-Gilad, J., Balcazar, J. L., Bibby, K., Elimelech, M., Weisbrod, N., Nir, O., Sued, O., Gillor, O., Alvarez, P. J., Cramer, S., Arnon, S., Walker, S., Yaron, S., Nguyen, T. H., Berchenko, Y., Hu, Y., Ronen, Z. & Bar-Zeev, E. (2020). Rethinking wastewater

- risks and monitoring in light of the COVID-19 pandemic. *Nature Sustainability*, 3, 981-990.
<https://doi.org/10.1038/s41893-020-00605-2>
- Boguniewicz_Zabłocka, J. & Cappodaglio, A. G. (2017). Sustainable wastewater treatment solutions for rural communities': public (centralized) or individual (on-site) - case study. *Economic and Environmental Studies*, 17(4:44), 1103-1119. <https://doi.org/10.25167/ees.2017.44.29>
- Boyjoo, Y., Pareek, V. K. & Ang, M. (2013). A review of greywater characteristics and treatment processes. *Water Science and Technology*, 67(7), 1403-1424.
<https://doi.org/10.2166/wst.2013.675>
- BSI. (2010). BSI Standards Publication BS 8525-1:2010 Greywater systems- Part 1: Code of practice, London: British Standards Institution.
- Buntat, Z., Iqbal, S. M. Z., Saburi, W. M. F. A., Adzis, Z., Sohaili, J. & Smith, I. R. (2015). Development of an integrated system for ozone treated harvested rainwater in perspective of green building scenario of Malaysia. *Journal of Environment and Earth Science*, 5(1), 51-60.
- Carrard, N., Foster, T. & Willetts, J. (2019). Groundwater as a source of drinking water in Southeast Asia and the Pacific: a multi-country review of current reliance and resource concerns. *Water*, 11(1605), 1-20. <https://doi.org/10.3390/w11081605>
- Cecconet, D., Callegari, A., Hlavínek, P. & Capodaglio, A. G. (2019). Membrane bioreactors for sustainable, fit-for-purpose greywater treatment: a critical review. *Clean Technologies and Environmental Policy*, 21(4), 745-762. <https://doi.org/10.1007/s10098-019-01679-z>
- Chen, J., Loeb, S. & Kim, J. H. (2017). LED revolution: Fundamentals and prospects for UV disinfection applications. *Environmental Science: Water Research & Technology*, 3(2), 188-202.
<https://doi.org/10.1039/C6EW00241B>
- Chirisa, I., Bandaiko, E., Matamanda, A., Mandisvika, G. (2017). Decentralized domestic wastewater systems in developing countries: the case study of Harare (Zimbabwe). *Applied Water Science*, 7, 1069-1078. <https://doi.org/10.1007/s13201-016-0377-4>

- Chrispim, M. C. & Nolasco, M. A. (2017). Greywater treatment using a moving bed biofilm reactor at a university campus in Brazil. *Journal of Cleaner Production*, 142, 290-296.
<https://doi.org/10.1016/j.jclepro.2016.07.162>
- Chukwuma, E. C., Ogbu, K. N. & Okonkwo, I. F. (2013). Quality assessment of direct harvested rainwater in parts of Anambra State, Nigeria. *International Journal of Agriculture and Biosciences*, 2(3), 120-123.
- Collivignarelli, M. C., Miino, M. C., Gomez, F. H., Torretta, V., Rada, E. C., & Sorlini, S. (2020). Horizontal flow constructed wetland for greywater treatment and reuse: an experimental case. *International Journal of Environmental Research and Public Health*, 17(7), 2317, 1-15.
<https://doi.org/10.3390/ijerph17072317>
- Cooke, F. M., Nordensvard, J., Saat G. B., Urban, F. & Siciliano, G. (2017). The limits of social protection: the case of hydropower dams and indigenous peoples' land. *Asia & the Pacific Policy Studies*, 4(3), 437-450. <https://doi.org/10.1002/app5.187>
- Corbould, C. (2013). Strategic Analysis Paper- Feeding the Cities: Is Urban Agriculture the Future of Food Security? Australia: Future Directions International Pty Ltd. International; Independent Strategic Analysis of Australia's Global Interests.
- Cotillas, S., Llanos, J., Moraleda, I., Cañizares, P. & Rodrigo, M. A. (2020). Scaling-up an integrated electrodisinfection-electrocoagulation process for wastewater reclamation. *Chemical Engineering Journal*, 380, 122415. <https://doi.org/10.1016/j.cej.2019.122415>
- D&D Ecotech Services. (2021). Groundwater RWH. [Online] Available at: <https://dndecotech.com/types-of-rainwater-harvesting/groundwater-rwh/> [Accessed: 13 August 2021].
- Dadonaite, B., Ritchie, H. & Roser, M. (2019). Diarrheal diseases. [Online] Available at: <https://ourworldindata.org/diarrheal-diseases> [Accessed: 8 August 2021].
- Deng, S., Xie, B., Kong, Q., Peng, S., Wang, H., Hu, Z., & Li, D. (2020). An oxic/anoxic-integrated and fe/c micro-electrolysis-mediated vertical constructed wetland for decentralized low-carbon

- greywater treatment. *Bioresource Technology*, 315(123802), 1-11.
<https://doi.org/10.1016/j.biortech.2020.123802>
- Disha, A. S., Harun, M. A., Akter, S., Billah, S. M., & Noman, M. A.-A. (2020). Reusing greywater for cultivation of capsicum frutescens and calendula officinalis. *Journal of Environmental Management*, 272(111088), 1-8. <https://doi.org/10.1016/j.jenvman.2020.111088>
- Dorji, U., Tenzin, U., Dorji, P., Pathak, N., Johir, M. A. H., Volpin, F., Dorji, C., Chernicharo, C. A. L., Tijing, L., Shon, H. & Phuntsho, S. (2021). Exploring shredded waste PET bottles as a biofilter media for improved on-site sanitation. *Process Safety and Environmental Protection*, 148, 370-381. <https://doi.org/10.1016/j.psep.2020.09.066>
- dos Santos, P. R. & Daniel, L. A. (2020). A review: organic matter and ammonia removal by biological activated carbon filtration for water and wastewater treatment. *International Journal of Environmental Science and Technology*, 17(1), 591-606. <https://doi.org/10.1007/s13762-019-02567-1>
- Doungmanee, P. (2016). The nexus of agricultural water use and economic development level. *Katsetsart Journal of Social Sciences*, 37(1), 38-45. <https://doi.org/10.1016/j.kjss.2016.01.008>
- Dwumfour-Asare, B., Nyarko, K. B., Essandoh, H. M. K. & Awuah, E. (2020). Domestic greywater flows and pollutant loads: a neighbourhood study within a university campus in Ghana: Characteristics of greywater in a residential neighbourhood. *Scientific African*, 9(e00489), 1-12. <https://doi.org/10.1016/j.sciaf.2020.e00489>
- Edwin, G. A., Gopalsamy, P. & Muthu, N. (2014). Characterization of domestic gray water from point source to determine the potential for urban residential reuse: a short review. *Applied Water Science*, 4, 39-49. <https://doi.org/10.1007/s13201-013-0128-8>
- Environmental Health. (2020). NSW Government Health. [Online] Available at: <https://www.health.nsw.gov.au/environment/water/Pages/rainwater.aspx> [Accessed: 13 August 2021].
- EPA. (2012). 2012 Guidelines for Water Reuse AR-1530: EPA/600/R-12/618. Washington, D.C.: United States Environmental Protection Agency.

- Eslami, H., Ehrampoush, M. H., Ghaneian, M. T., Mokhtari, M. & Ebrahimi, A. (2017). Effect of Organic Loading Rates on biodegradation of linear alkyl benzene sulfonate, oil and grease in greywater by Integrated Fixed-film Activated Sludge (IFAS). *Journal of Environmental Management*, 193, 312-317. <https://doi.org/10.1016/j.jenvman.2017.02.038>
- Estévez, S., González-García, S., Feijoo, G. & Moreira, M. T. (2021). How decentralized treatment can contribute to the symbiosis between environmental protection and resource recovery. *Science of the Total Environment*, 1-13. <https://doi.org/10.1016/j.scitotenv.2021.151485>
- Expert Market Research. (2020). Report: Global Rainwater Harvesting Market Outlook. [Online] Available at: <https://www.expertmarketresearch.com/reports/rainwater-harvesting-market> [Accessed: 13 August 2021].
- Ezugbe, E. O. & Rathilal, S. (2020). Membrane technologies in wastewater treatment: a review. *Membranes*, 10(5), 89, 1-28. <https://doi.org/10.3390/membranes10050089>
- Falsanisi, D., Liberti, L. & Notarnicola, M. (2010). Ultrafiltration (UF) pilot plant for municipal wastewater reuse in agriculture: Impact of the operation mode on process performance. *Water*, 2(4), 872-885. <https://doi.org/10.3390/w2040872>
- Formisano, F., Fiorentino, A., Rizzo, L., Carotenuto, M., Pucci, L., Giugni, M. & Lofrano, G. (2016). Inactivation of Escherichia coli and Enterococci in urban wastewater by sunlight/PAA and sunlight/H2O2 processes. *Process Safety and Environmental Protection*, 104, 178-184. <https://doi.org/10.1016/j.psep.2016.09.003>
- Foster, S., Chilton, J., Moench, M., Cardy, F., & Schiffler, M. (2000). Groundwater in Rural Development: Facing the Challenges of Supply and Resource Sustainability; World Bank Technical Paper No. 463. Washington, D.C.: The International Bank for Reconstruction and Development/ THE WORLD BANK.
- Germec, M., Demirci, A. & Turhan, I. (2020). Biofilm reactors for value-added products production: An in-depth review. *Biocatalysis and Agricultural Biotechnology*, 27(101662). <https://doi.org/10.1016/j.bcab.2020.101662>

- Ghaitidak, D. M. & Yadav, K. D. (2015). Effect of coagulant in greywater treatment for reuse: selection of optimal coagulation condition using Analytic Hierarchy Process. *Desalination and Water Treatment*, 55(4), 913-925. <https://doi.org/10.1080/19443994.2014.924036>
- Goncalves, R. F., Vaz, L. d. O., Peres, M. & Merlo, S. S. (2021). Microbiological risk from non-potable reuse of greywater treated by anaerobic filters associated to vertical constructed wetlands. *Journal of Water Process Engineering*, 39(101751), 1-11. <https://doi.org/10.1016/j.jwpe.2020.101751>
- Graf. (2021). Wastewater treatment: Tanks underground. [Online] Available at: <https://www.graf-water.com/download/catalogues/wastewater-treatment.html> [Accessed: 13 August 2021].
- Greyter, Inc. (2021a). Multi Unit Residential and Commercial Buildings. [Online] Available at: <https://greyter.com/commercial/> [Accessed: 13 August 2021].
- Greyter, Inc. (2021b). Residential Homes: THE GREYTER HOME TM RESIDENTIAL WATER RECYCLING SYSTEM. [Online] Available at: <https://greyter.com/residential/> [Accessed: 13 August 2021].
- Gun, J. V. D. (2021). Chapter 24 - Groundwater resources sustainability, Editor(s): Abhijit Mukherjee, Bridget R. Scanlon, Alice Aureli, Simon Langan, Huaming Guo, Andrew A. McKenzie, Global Groundwater, Elsevier, pp. 331-345. <https://doi.org/10.1016/B978-0-12-818172-0.00024-4>
- Halalsheh, M., Dalahmeh, S., Sayed, M., Suleiman, W., Shareef, M., Mansour, M., & Safi, M. (2008). Grey water characteristics and treatment options for rural areas in Jordan. *Bioresource Technology*, 99(14), 6635-6641. <https://doi.org/10.1016/j.biortech.2007.12.029>
- Hammed, T. B., Sridhar, M.K.C. & Fawole, O. B. (2017). A low cost 'umbrella rainwater harvester' for open air markets in Nigeria: design, fabrication, yield and quality assessment. *International Journal of Scientific & Engineering Research*, 8(7), 1451-1458.
- Hassaballah, A. H., Bhatt, T., Nyitrai, J., Dai, N. & Sassoubre, L. (2020). Inactivation of E. coli, Enterococcus spp., somatic coliphage, and Cryptosporidium parvum in wastewater by peracetic acid (PAA), sodium hypochlorite, and combined PAA-ultraviolet disinfection. *Environmental Science: Water Research & Technology*, 6(1), 197-209. <https://doi.org/10.1039/C9EW00837C>

- Heritage Tanks. (2016). Evarain Circular Tank Gutter System. [Online] Available at: <https://heritagetanks.com.au/gutter-technology/> [Accessed: 13 August 2021].
- Hinkebein, T. E. (2004). Desalination: Limitations and Challenges. In: P. Norling, F. Wood-Black & T. Masciangioli, eds. *Water and Sustainable Development: Opportunities for the Chemical Sciences: A Workshop Report to the Chemical Sciences Roundtable*. Washington, D.C.: National Academies Press, pp. 29-39.
- Hitchner, B. (2016). Nexus E Water- "Onsite Water Reuse" for the Home. [Online] Available at: <https://ccag.ca.gov/wp-content/uploads/2016/05/Nexus-eWater-CCAG.pdf> [Accessed: 13 August 2021].
- Hoang, T. T. L., Vigneswaran, S., Ngo, H. H., Kandasamy, J., Shim, W. G., Chaudhary, D. S., Gotety, P. & Peiris, P. (2008). Performance evaluation and mathematical modelling of granular activated carbon biofiltration in wastewater treatment. *Korean Journal of Chemical Engineering*, 25(2), 259-267. <https://doi.org/10.1007/s11814-008-0046-x>
- Hoang, T., Shon, H., Chaudhary, D., Vigneswaran, S. & Ngo, H. (2004). Granular activated carbon (GAC) biofilter for low strength wastewater treatment. *Fluid/Particle Separation Journal*, 16(2), 185-191.
- Hocaoglu, S. M., Insel, G., Cokgor, E. U., Baban, A., & Orhon, D. (2010). Cod fractionation and biodegradation kinetics of segregated domestic wastewater: black and grey water fractions. *Journal of Chemical Technology & Biotechnology*, 85(9), 1241-1249. <https://doi.org/10.1002/jctb.2423>
- Hourlier, F., Massé, A., Jaouen, P., Lakel, A., Gérente, C., Faur, C. & Cloirec, P. L. (2010). Membrane process treatment for greywater recycling: investigations on direct tubular nanofiltration. *Water Science & Technology*, 62(7), 1544-1550. <https://doi.org/10.2166/wst.2010.435>
- Huang, Y., Li, P., Li, H., Zhang, B. & He, Y. (2021). To centralize or decentralize? A systematic framework for optimizing rural wastewater treatment planning. *Journal of Environmental Management*, 300(113673), 1-10. <https://doi.org/10.1016/j.envman.2021.113673>

- Igoud, S., Souahi, F., Chitour, C. E., Amrouche, L., Lamaa, C., Chekir, N. & Chouikh, A. (2015). Wastewater disinfection using ultraviolet (UVA, UVC) and solar radiation. *Desalination and Water Treatment*, 56(10), 2646-2652. <https://doi.org/10.1080/19443994.2014.968909>
- Innovative Water Solutions. (2020). Residential Rainwater System Installation. [Online] Available at: <https://www.watercache.com/rainwater/residential> [Accessed: 13 August 2021].
- International Water Association. (2018). Wastewater Report 2018: The Reuse Opportunity. London: International Water Association.
- Islam, M. M., Chou, F. N. F., Kabir, M. R. & Liaw, C. H. (2010). Rainwater: a potential alternative source for scarce safe drinking and arsenic contaminated water in Bangladesh. *Water Resources Management*, 24, 3987-4008. <https://doi.org/10.1007/s11269-010-9643-7>
- Jabornig, S. & Podmirseg, S. M. (2015). A novel fixed fibre biofilm membrane process for on-site greywater reclamation requiring no fouling control. *Biotechnology and Bioengineering*, 112(3), 484-493. <https://doi.org/10.1002/bit.25449>
- Jabri, K. M., Fiedler, T., Saidi, A., Nolde, E., Ogurek, M., Geissen, S.-U., & Bousselmi, L. (2019). Steady-state modeling of the biodegradation performance of a multistage moving bed biofilm reactor (mbbr) used for on-site greywater treatment. *Environmental Science and Pollution Research*, 26, 19047-19062. <https://doi.org/10.1007/s11356-018-3984-9>
- Jamrah, A., Al-Omari, A., Al-Qasem, L. & Abdel, N. (2006). Assessment of availability and characteristics of greywater in Amman. *Water International*, 31(2), 210-220. <https://doi.org/10.1080/02508060.2006.9709671>
- Kahinda, J. M. & Taigbenu, A. E. (2011). Rainwater harvesting in South Africa: challenges and opportunities. *Physics and Chemistry of the Earth*, 36, 968-976. <https://doi.org/10.1016/j.pce.2011.08.011>
- Kalkan, C., Yapsakli, K., Mertoglu, B., Tufan, D. & Saatci, A. (2011). Evaluation of biological activated carbon (BAC) process in wastewater treatment secondary effluent for reclamation purposes. *Desalination*, 265(1-3), 266-273. <https://doi.org/10.1016/j.desal.2010.07.060>

- Karpova, T., Pekonen, P., Gramstad, R., Öjstedt, U., Laborda, S., Heinonen-Tanski, H., Chávez, A. & Jiménez, B. (2013). Performic acid for advanced wastewater disinfection. *Water Science & Technology*, 68(9), 2090-2096. <https://doi.org/10.2166/wst.2013.468>
- Kayhanian, M., Fruchtmann, B. D., Gulliver, J. S., Montanaro, C., Ranieri, E. & Wuertz, S. (2012). Review of highway runoff characteristics: comparative analysis and universal implications. *Water Research*, 46(20), 6609-6624. <https://doi.org/10.1016/j.watres.2012.07.026>
- Khayan, K., Heru Husodo, A., Astuti, I., Sudarmadji, S., & Sugandawaty Djohan, T. (2019). Rainwater as a source of drinking water: health impacts and rainwater treatment. *Journal of Environmental and Public Health*, 1760950. <https://doi.org/10.1155/2019/1760950>
- Khosravanipour Mostafazadeh, A., Benguit, A. T., Carabin, A., Drogui, P. & Brien, E. (2019). Development of combined membrane filtration, electrochemical technologies, and adsorption processes for treatment and reuse of laundry wastewater and removal of nonylphenol ethoxylates as surfactants. *Journal of Water Process Engineering*, 28, 277-292. <https://doi.org/10.1016/j.jwpe.2019.02.014>
- Khuntia, H. K., Chandrashekar, S. & Chanakya, H. N. (2019). Treatment of household greywater laden with household chemical products in a multi-chambered anaerobic biofilm reactor. *Sustainable Cities and Society*, 51(101783). <https://doi.org/10.1016/j.scs.2019.101783>
- Kumar, R., Singh, R. D. & Sharma, K. D. (2005). Water resources of India. *Current Science*, 89(5), 794-811. <http://www.jstor.org/stable/24111024>
- Laminack, K. D. (2014). Green roof water harvesting and recycling effects on soil and water chemistry and plant physiology. *Master's Thesis*, Texas A&M University, United States.
- Lani, N. H. M., Yusop, Z. & Syafiuddin, A. (2018). A review of rainwater harvesting in Malaysia: prospects and challenges. *Water*, 10(506), 1-21. <https://doi.org/10.3390/w10040506>
- La Rosa, G., Iaconelli, M., Mancini, P., Bonanno Ferraro, G., Veneri, C., Bonadonna, L., Lucentini, L. & Suffredini, E. (2020). First detection of SARS-CoV-2 in untreated wastewaters in Italy. *Science of the Total Environment*, 736(139652). <https://doi.org/10.1016/j.scitotenv.2020.139652>

- 1
2
3
4 Leal, L. H., Temmink, H., Zeeman, G. & Buisman, C. J. N. (2010). Bioflocculation of grey water for
5
6 improved energy recovery within decentralized sanitation concepts. *Bioresource & Technology*,
7
8 101(23), 9065-9070. <https://doi.org/10.1016/j.biortech.2010.07.047>
9
- 10 Lee, M., Kim, M., Kim, Y. & Han, M. (2017) Consideration of rainwater quality parameters for drinking
11
12 purposes: A case study in rural Vietnam. *Journal of Environmental Management*, 200, 400-406.
13
14 <https://doi.org/10.1016/j.jenvman.2017.05.072>
15
- 16 Leonard, M., Gilpin, B., Robson, B. & Wall, K. (2016). Field of study of the composition of greywater
17
18 and comparison of microbiological indicators of water quality in on-site systems. *Environmental*
19
20 *Monitoring and Assessment*, 188(8):475, 1-16. <https://doi.org/10.1007/s10661-016-5442-9>
21
22
- 23 Leong, C. & Lebel, L. (2020). Can conformity overcome the yuck factor? explaining the choice for
24
25 recycled drinking water. *Journal of Cleaner Production*, 242(118196), 1-7.
26
27 <https://doi.org/10.1016/j.jclepro.2019.118196>
28
- 29 Li, Z., Boyle, F. & Reynolds, A. (2010). Rainwater harvesting and greywater treatment systems for
30
31 domestic application in Ireland. *Desalination*, 260(1-3), 1-8.
32
33 <https://doi.org/10.1016/j.desal.2010.05.035>
34
- 35 Liberti, L., Lopez, A, Notarnicola, M., Barnea, N., Pedahzur, R. & Fattal, B. (2000). Comparison of
36
37 advanced disinfecting methods for municipal wastewater reuse in agriculture. *Water Science &*
38
39 *Technology*, 42(1-2), 215-220. <https://doi.org/10.2166/wst.2000.0316>
40
41
- 42 Liyandeniya, A.B., Deeyamulla, M.P. & Priyantha, N. (2020). Source apportionment of rainwater
43
44 chemical composition in wet precipitation at Kelaniya in Sri Lanka. *Air Quality Atmosphere &*
45
46 *Health*, 13, 1497–1504. <https://doi.org/10.1007/s11869-020-00903-w>
47
48
- 49 Maniam, G., Poh, P. E., Htar, T. T., Poon, W. C., & Chuah, L. H. (2021). Water Literacy in the
50
51 Southeast Asian Context: Are We There Yet? *Water*, 13(16), 2311.
52
53 <https://doi.org/10.3390/w13162311>
54
- 55 Manouchehri, M. & Kargari, A. (2017). Water recovery from laundry wastewater by the cross flow
56
57 microfiltration process: A strategy for water recycling in residential buildings. *Journal of Cleaner*
58
59 *Production*, 168, 227-238. <https://doi.org/10.1016/j.jclepro.2017.08.211>
60

- Matala Water Technology Co., Ltd. (2011). Matala: Make water alive. [Online] Available at:
https://www.matala.com.tw/upload_files/ebook/wastewater/Matala_Filter_media-Wastewater.pdf
[Accessed: 13 August 2021].
- McCarthy, D.T., Hathaway, J.M., Hunt, W.F. & Deletic, A. (2012). Intra-event variability of *Escherichia coli* and total suspended solids in urban stormwater runoff. *Water Research*, 46(20), 6661-6670.
<https://doi.org/10.1016/j.watres.2012.01.006>
- Mazuki, N. I. M., Teow, Y. H., Ho, K. C. & Mohammad, A.W. (2020). Techno-economic analysis of single disinfection units and integrated disinfection systems for sewage effluent reclamation. *Journal of Water Process Engineering*, 36(101398). <https://doi.org/10.1016/j.jwpe.2020.101398>
- Momeni, M., Behzadian, K., Yousefi, H. & Zahedi, S. (2021). A Scenario-Based Management of Water Resources and Supply Systems Using a Combined System Dynamics and Compromise Programming Approach. *Water Resources Management*, 35, 4233-4250.
<https://doi.org/10.1007/s11269-021-02942-z>
- Morel, A. & Diener, S. (2006). Greywater Management in Low and Middle-Income Countries: Review of different treatment systems for households or neighbourhoods. Dubendorf: Swiss Federal Institute of Aquatic Science and Technology (Eawag).
- Moreno-Andrés, J., Rueda-Márquez, J. J., Homola, T., Vielma, J., Moríñigo, M. Á., Mikola, A., Sillanpää, M., Acevedo-Merino, A., Nebot, E. & Levchuk, I. (2020). A comparison of photolytic, photochemical and photocatalytic processes for disinfection of recirculation aquaculture systems (RAS) streams. *Water Research*, 181(115928).
<https://doi.org/10.1016/j.watres.2020.115928>
- Morley, M., Behzadian, K., Kapelan, Z. & Ugarelli, R. (2016). Decision support system for metabolism-based transition to urban water systems of tomorrow. *Water Science and Technology: Water Supply*, 16(3), 855-863. <https://doi.org/10.2166/ws.2016.007>
- Mostafa, M. G. & Shafiuzzaman, S. M. (2008). Potential use of monsoon rainwater for drinking purpose in Bangladesh. *Water Practice & Technology*, 3(3), 1-8.
<https://doi.org/10.2166/WPT.2008068>

- Musingafi, M. C. C. (2013). A global overview of potable water resources availability and accessibility in Southern Africa. *Journal of Natural Sciences*, 3(2), 26-34.
<https://doi.org/10.13140/2.1.1163.1041>
- National Geographic. (2017). National Geographic. [Online] Available at:
<https://www.nationalgeographic.org/article/all-about-climate/> [Accessed 11 May 2021].
- Niekerk, A. M. V. & Schneider, B. (2013). Implementation Plan for Direct and Indirect Water Re-use for Domestic Purposes - Sector Discussion Document (WRC Project No. KV 320/13). Gezina, South Africa: Water Research Commission.
- Noutsopoulos, C., Andreadakis, A., Kouris, N., Charchousi, D., Mendrinou, P., Galani, A., . . . Koumaki, E. (2018). Greywater characterization and loadings - physicochemical treatment to promote onsite reuse. *Journal of Environmental Management*, 216, 337-346.
<https://doi.org/10.1016/j.jenvman.2017.05.094>
- OECD/Eurostat. (2018). Data Collection Manual for the OECD/Eurostat Joint Questionnaire on Inland Waters and Eurostat regional water questionnaire: Concepts, definitions, current practices, evaluations and recommendations Version 4. Luxembourg: Eurostat.
- Oh, K. S., Leong, J. Y. C., Poh, P. E., Chong, M. N. & Lau, E. V. (2018). A review of greywater recycling related issues: challenges and future prospects in Malaysia. *Journal of Cleaner Production*, 171, 17-29. <https://doi.org/10.1016/j.jclepro.2017.09.267>
- Oktor, K. & Celik, D. (2019). Treatment of wash basin and bathroom greywater with chlorella variabilis and reusability. *Journal of Water Process Engineering*, 31(100857), 1-7.
<https://doi.org/10.1016/j.jwpe.2019.100857>
- Oladoja, N. A. (2017). Appropriate technology for domestic wastewater management in under-resourced regions of the world. *Applied Water Science*, 7, 3391-3406.
<https://doi.org/10.1007/s13201-016-0495-z>
- Omar, K., Aziz, N., Amr, S. & Palaniandy, P. (2017). Removal of lindane and Escherichia coli (E. coli) from rainwater using photocatalytic and adsorption treatment processes. *Global Nest Journal*, 19, 191-198.

- Oteng-Peprah, M., de Vries, N. K. & Acheampong, M. A. (2018). Greywater characterization and generation rates in a peri urban municipality of a developing country. *Journal of Environmental Management*, 206, 498-506. <https://doi.org/10.1016/j.jenvman.2017.10.068>
- Pablos, C., van Grieken, R., Marugán, J. & Muñoz, A. (2012). Simultaneous photocatalytic oxidation of pharmaceuticals and inactivation of *Escherichia coli* in wastewater treatment plant effluents with suspended and immobilised TiO₂. *Water Science & Technology*, 65(11), 2016-2023. <https://doi.org/10.2166/wst.2012.868>
- Pala, G. K., Pathivada, A. P., Velugoti, S. J. H., Yerramsetti, C., Veeranki, S. (2021). Rainwater harvesting- a review on conservation, creation & cost-effectiveness. *Materials Today: Proceedings*, 45, 6567-6571. <https://doi.org/10.1016/j.matpr.2020.11.593>
- Palmarin, M. J. & Young, S. (2019). Comparison of the treatment performance of a hybrid and conventional membrane bioreactor for greywater reclamation. *Journal of Water Process Engineering*, 28, 54-59. <https://doi.org/10.1016/j.jwpe.2018.12.012>
- Perdana, M. C., Hadisusanto, S. & Purnama, I. L. (2020). Implementation of a full-scale constructed wetland to treat greywater from tourism in Suluban Uluwatu Beach, Bali, Indonesia. *Heliyon*, 6(0), e05038, 1-9. <https://doi.org/10.1016/j.heliyon.2020.e05038>
- Peter-Varbanets, M., Zurbrugg, C., Swartz, C. & Pronk, W. (2009). Decentralized systems for potable water and the potential of membrane technology. *Water Research*, 43, 245-265. <https://doi.org/10.1016/j.watres.2008.10.030>
- Piadeh, F., Ahmadi, M. & Behzadian, K., (2018). Reliability assessment for hybrid systems of advanced treatment units of industrial wastewater reuse using combined event tree and fuzzy fault tree analyses. *Journal of Cleaner Production*, 201, 958-973. <https://doi.org/10.1016/j.jclepro.2018.08.052>
- Pramanik, B. K., Roddick, F. A. & Fan, K. (2014). Effect of biological activated carbon pre-treatment to control organic fouling in the microfiltration of biologically treated secondary effluent. *Water Research*, 63, 147-157. <https://doi.org/10.1016/j.watres.2014.06.014>

- Prescott, M. F., Dobbie, M. F. & Ramirez-Lovering, D. (2021). Green infrastructure for sanitation in settlements in the Global South: a narrative review of socio-technical systems. *Sustainability*, 13(2071), 1-21. <https://doi.org/10.3390/su13042071>
- Pronk, W., Ding, A., Morgenroth, E., Derlon, N., Desmond, P., Burkhardt, M., Wu, B. & Fane, A. G. (2019). Gravity-driven membrane filtration for water and wastewater treatment: A review. *Water Research*, 149, 553-565. <https://doi.org/10.1016/j.watres.2018.11.062>
- Pure Water LLC. (2010). Grey Water Treatment Skid System. [Online] Available at: <http://purewater2000.com/greywatermanual.html> [Accessed: 13 August 2021].
- Putri, P. W. (2017). A decentralised approach to wastewater management in the urbanising region: the case of Jakarta, Indonesia. *Urbanisation*, 2(2), 83-97. <https://doi.org/10.1177/2455747117740439>
- Radaideh, J., Al-Zboon, K., Al-Harabsheh, A., Al-Adamat, R. (2009). Quality assessment of harvested rainwater for domestic uses. *Jordan Journal of Earth and Environmental Sciences*, 2(1), 26-31.
- Rainwater Harvesting. (2015). What Do We Do? [Online] Available at: <http://www.rainyfilters.com> [Accessed: 13 August 2021].
- Randazzo, W., Cuevas-Ferrando, E., Sanjuán, R., Domingo-Calap, P. & Sánchez, G. (2020). Metropolitan wastewater analysis for COVID-19 epidemiological surveillance. *International Journal of Hygiene and Environmental Health*, 230(113621). <https://doi.org/10.1016/j.ijheh.2020.113621>
- Raphael, D. O., Okunade, D. A., Ogedengbe, K. & Adekunle, O. A. (2020). Assessment of a batch-flow free water surface constructed wetland planted with *rhynchospora corymbosa* (L.) britton for campus greywater treatment. *Environmental Science and Pollution Research*, 27(4), 4275-4283. <https://doi.org/10.1007/s11356-019-07095-6>
- Reungoat, J., Escher, B., Macova, M. & Keller, J. (2011). Biofiltration of wastewater treatment plant effluent: Effective removal of pharmaceuticals and personal care products and reduction of toxicity. *Water Research*, 45(9), 2751-2762. <https://doi.org/10.1016/j.watres.2011.02.013>

- Reungoat, J., Escher, B., Macova, M., Argaud, F., Gernjak, W. & Keller, J. (2012). Ozonation and biological activated carbon filtration of wastewater treatment plant effluents. *Water Research*, 46(3), 863-872. <https://doi.org/10.1016/j.watres.2011.11.064>
- Reynaert, E., Hess, A. & Morgenroth, E. (2021). Making waves: why water reuse frameworks need to co-evolve with emerging small-scale technologies. *Water Research X*, 11(100094), 1-5. <https://doi.org/10.1016/j.wroa.2021.100094>
- Rhim, J. A. (2006). Characteristics of adsorption and biodegradation of dissolved organic carbon in biological activated carbon pilot plant. *Korean Journal of Chemical Engineering*, 23(1), 38-42. <https://doi.org/10.1007/BF02705690>
- Ritchie, H. & Roser, M. (2017). Water Use and Stress. [Online] Available at: <https://ourworldindata.org/water-use-stress> [Accessed: 13 August 2021].
- Ritchie, H. & Roser, M. (2021a). Clean Water. [Online] Available at: <https://ourworldindata.org/water-access> [Accessed: 13 August 2021].
- Ritchie, H. & Roser, M. (2021b). Sanitation. [Online] Available at: <https://ourworldindata.org/sanitation#access-to-safe-sanitation> [Accessed: 13 August 2021].
- Rodrigues, S. (2021). China Reclaimed Water Reuse Regulations. [Online] Available at: https://www.iwa-network.org/filemanager-uploads/WQ_Compendium/Cases/China%20Reclaimed.pdf [Accessed: 11 May 2021].
- Rosegrant, M. W., Cai, X. & Cline, S. A. (2002). Food Policy Report: Global Water Outlook to 2025-Averting an Impending Crisis. Washington, DC: International Food Policy Research Institute.
- Saidi, A., Masmoudi, K., Nolde, E., El Amrani, B. & Amraoui, F. (2017). Organic matter degradation in a greywater recycling system using a multistage moving bed biofilm reactor (MBBR). *Water Science & Technology*, 76(12), 3328-3339. <https://doi.org/10.2166/wst.2017.499>
- Salehi, M., Aghilinasrollahabadi, K. & Esfandarani M. S. (2020). An investigation of stormwater quality variation within an industry sector using the self-reported data collected under the stormwater monitoring program. *Water*. 12(11):3185. <https://doi.org/10.3390/w12113185>

- Samuel, M. P., Senthilvel, S. & Mathew, A. C. (2012). Water quality assessment of various forms of rainwater and statistical studies on physico-chemical characteristics of stormwater in Coimbatore, India. *Nature Environment and Pollution Technology*, 11(1), 23-28.
- Sato, T., Qadir, M., Yamamoto, S., Endo, T., & Zahoor, A. (2013). Global, regional, and country level need for data on wastewater generation, treatment, and use. *Agricultural Water Management*, 130, 1-13. <https://doi.org/10.1016/j.agwat.2013.08.007>
- Sha'arani, S., Azizan, S. N. F., Akhir, F. N. M., Yuzir, M. A. M., Othman, N. a., Zakaria, Z., . . . Hara, H. (2019). Removal efficiency of Gram-positive and Gram-negative bacteria using a natural coagulant during coagulation, flocculation, and sedimentation processes. *Water Science and Technology*, 80(9), 1787-1795. <https://doi.org/10.2166/wst.2019.43>
- Shaikh, I. N. & Ahammed, M. M. (2020). Quantity and *quality* characteristics of greywater: a review. *Journal of Environmental Management*, 261(110266), 1-15. <https://doi.org/10.1016/j.jenvman.2020.110266>
- Sharaf, A. & Liu, Y. (2021). Mechanisms and kinetics of greywater treatment using biologically active granular activated carbon. *Chemosphere*, 263(128113). <https://doi.org/10.1016/j.chemosphere.2020.128113>
- Sievers, J. C. & Londong, J. (2018). Characterization of domestic graywater and graywater solids. *Water Science & Technology*, 77.5, 1196-1203. <https://doi.org/10.2166/wst.2017.627>
- Sigmon, C., Shin, G. A., Mieog, J. & Linden, K. G. (2015). Establishing Surrogate - Virus Relationships for Ozone Disinfection of Wastewater. *Environmental Engineering Science*, 32(6), 451-460. <https://doi.org/10.1089/ees.2014.0496>
- Sivakumar, M. & Ramezaniapour, M. (2012). Green Technology Options for Greywater Treatment. Mangalore, s.n., pp. 1-18.
- Solaripedia. Project Solar Trees in City Gardens (Singapore). 2014. Available online: https://www.solaripedia.com/13/416/6629/solar_supertrees_cellular_paving_photo.html (accessed on 28 January 2022).

- Stefanakis, A. I. (2020). Constructed wetlands for sustainable wastewater treatment in hot and arid climates: opportunities, challenges and case studies in the Middle East. *Water*, 12(1665), 1-23. <https://doi.org/10.3390/w12061665>
- Stormsaver. (2021). Shallow Dig (SD) Tank Product Data Sheet. [Online] Available at: <https://www.stormsaver.com/content/documents/Monsoon%20Garden%20Package%20Data%20sheet%20-%202020.pdf> [Accessed: 13 August 2021].
- Subramanian, G. P., Raj, A. V., Jamwal, P., Connelly, S., Yeluripati, J., Richards, S., Ellis, R., Rao, L. (2020). Decentralized treatment and recycling of greywater from a school in rural India. *Journal of Water Process Engineering*, 38(1010695), 1-13. <https://doi.org/10.1016/j.jwpe.2020.101695>
- Tatlock, C. W. (2006). Council on Foreign Relations. [Online] Available at: <https://www.cfr.org/backgrounder/water-stress-sub-saharan-africa> [Accessed: 13 August 2021].
- Tombola, R., Buttiglieri, G., Auset, M. & Gonzalez-Olmos, R. (2019). Recycled corrugated wire hose cover as biological carriers for greywater treatment in a sequential batch biofilm reactor. *Journal of Environmental Management*, 240, 475-484. <https://doi.org/10.1016/j.jenvman.2019.02.116>
- Tortajada, C. (2020). Contributions of recycled wastewater to clean water and sanitation sustainable development goals. *npj Clean Water*, 3(22), 1-6. <https://doi.org/10.1038/s41545-020-0069-3>
- Truu, M., Oopkaup, K., Krustok, I., Koiv-Vainik, M., Nolvak, H., & Truu, J. (2019). Bacterial community activity and dynamics in the biofilm of an experimental hybrid wetland system treating greywater. *Environmental Science and Pollution Research*, 26, 4013-4026. <https://doi.org/10.1007/s11356-018-3940-8>
- U.S. Department of Health & Human Services. (2009). Centers for Disease Control and Prevention. [Online] Available at: https://www.cdc.gov/healthywater/drinking/public/water_sources.html#one [Accessed: 13 August 2021].
- Uceвли, O. & Kaya, Y. (2021). A comparative study of membrane filtration, electrocoagulation, chemical coagulation and their hybrid processes for greywater treatment. *Journal of Environmental Chemical Engineering*, 9(1):104946), 1-14. <https://doi.org/10.1016/j.jece.2020.104946>

- UN. (2019). The Sustainable Development Goals Report 2019. New York: United Nations.
- UN. (2020). The Sustainable Development Goals Report 2020. New York: United Nations.
- UN. (2021). The Sustainable Development Goals Report 2021. New York: United Nations.
- UN Water. (2013). Water Security & the Global Water Agenda: A UN-Water Analytical Brief. Ontario: United Nations University- Institute for Water, Environment & Health.
- UN Water/Africa. (2004). The Africa Water Vision for 2025: Equitable and Sustainable Use of Water for Socioeconomic Development. Addis Ababa: Economic Commission for Africa.
- UN, Department of Economic and Social Affairs, Population Division. (2019). World Population Prospects 2019: Highlights. New York: United Nations.
- UNEP. (2014) Water and Energy Efficiency: Information Brief. Nairobi, Kenya: United Nations Environment Programme.
- UNESCO. (2018). Policy Briefs on Internal Migration in Southeast Asia. Bangkok, Thailand: United Nations Educational, Scientific and Cultural Organization. Available at: <https://bangkok.unesco.org/index.php/content/policy-briefs-internal-migration-southeast-asia> [Accessed: 30 May 2021].
- UNESCO. (2021). The United Nations World Water Development Report 2021: Valuing Water. Paris: United Nations Educational, Scientific and Cultural Organization.
- UNESCO World Water Assessment Programme. (2017). The United Nations world water development report, 2017: Wastewater: the untapped resource. Paris: United Nations Educational, Scientific and Cultural Organization.
- UNESCO World Water Assessment Programme. (2020). The United Nations World Water Development Report 2020: Water and Climate Change. Paris: United Nations Educational, Scientific and Cultural Organization.
- UNICEF & WHO. (2019). Progress on household drinking water, sanitation and hygiene 2000-2017: Special focus on inequalities. New York: United Nations Children's Fund (UNICEF) and World Health Organization.

- UN-Water. (2021). Summary Progress Update 2021 – SDG 6 – water and sanitation for all. Geneva, Switzerland: United Nations Water.
- USEPA. (2010). Long Term 2 Enhanced Surface Water Treatment Rule Toolbox Guidance Manual.
- Water Field Technology. (2021). Dual Intensity Online Filter- Rain Water Harvesting and Ground Water Management. [Online] Available at: <https://waterfieldindia.com/products/rainy-filter/> [Accessed: 13 August 2021].
- Water UK. (2020). Water UK. [Online] Available at: <https://www.water.org.uk/advice-for-customers/water-and-health/> [Accessed: 13 August 2021].
- Watts. (2017). Rainwater. Engineered. Leading The Way With Sustainable Water Solutions. [Online] Available at: <https://www.watts.com/dfsmedia/0533dbba17714b1ab581ab07a4cbb521/25784-source/raincycle-rainwater-harvesting-system> [Accessed: 13 August 2021].
- WHO. (2017). Diarrhoeal disease: Key facts. [Online] Available at: <https://www.who.int/news-room/fact-sheets/detail/diarrhoeal-disease> [Accessed: 13 August 2021].
- WISY. (2020). Application: Our Filtration Technique is Unique. [Online] Available at: <http://www.wisysgreentech.com/application.php> [Accessed: 13 August 2021].
- Wölfel, R., Corman, V. M., Guggemos, W., Seilmaier, M., Zange, S., Müller, M. A., Niemeyer, D., Jones, T. C., Vollmar, P., Rothe, C., Hoelscher, M., Bleicker, T., Brünink, S., Schneider, J., Ehmann, R., Zwirgmaier, K., Drosten, C. & Wendtner, C. (2020). Virological assessment of hospitalized patients with COVID-2019. *Nature*, 581(7809), 465-469. <https://doi.org/10.1038/s41586-020-2196-x>
- World Health Organization. (2018). A Global Overview of National Regulations and Standards for Drinking-water Quality. [Online] Available at: <https://apps.who.int/iris/bitstream/handle/10665/272345/9789241513760-eng.pdf?ua=1> [Accessed: 16 August 2021].
- Zabidi, H. A., Goh, H. W., Chang, C. K., Chan, N. W., & Zakaria, N. A. (2020). A review of roof and pond rainwater harvesting systems for water security: the design, performance and way forward. *Water*, 12(11), 3163, 1-22. <https://doi.org/10.3390/w12113163>

- Zavala, M. A. L., Vega, R. C. & Miranda, R. A. L. (2016). Potential of rainwater harvesting and greywater reuse for water consumption reduction and wastewater minimization. *Water*, 8(6), 264, 1-18. <https://doi.org/10.3390/w8060264>
- Zhang, X., He, J., Lei, Y., Qiu, Z., Cheng, S. & Yang, X. (2019). Combining solar irradiation with chlorination enhances the photochemical decomposition of microcystin-LR. *Water Research*, 159, 324-332. <https://doi.org/10.1016/j.watres.2019.05.030>
- Zheng, T., Li, W., Ma, Y., Liu, J. & Ren, J. (2020). Greywater: Understanding biofilm bacteria succession, pollutant removal and low sulfide generation in small diameter gravity sewers. *Journal of Cleaner Production*, 268(122426). <https://doi.org/10.1016/j.jclepro.2020.122426>
- Zhou, Y., Guo, B., Li, R., Zhang, L., Xia, S. & Liu, Y. (2020). Treatment of grey water (GW) with high linear alkylbenzene sulfonates (LAS) content and carbon/nitrogen (C/N) ratio in an oxygen-based membrane biofilm reactor (O2-MBfR). *Chemosphere*, 258(127363). <https://doi.org/10.1016/j.chemosphere.2020.127363>

Further Reading

- Ormerod, K.J. (2016). Illuminating elimination: public perception and the production of potable water reuse. *WIREs Water*, 3, 537-547. <https://doi.org/10.1002/wat2.1149>
- Arora, M., Malano, H., Davidson, B., Nelson, R. & George, B. (2015). Interactions between centralized and decentralized water systems in urban context: A review. *WIREs Water*, 2, 623-634. <https://doi.org/10.1002/wat2.1099>
- Nazif, S., Tavakolifar, H., Abbasizadeh, H. & Eslamian, S. (2021). Comparing Rainwater Storage Options. In *Handbook of Water Harvesting and Conservation*, S. Eslamian (Ed.). <https://doi.org/10.1002/9781119478911.ch25>

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

Pradhan, S., Al-Ghamdi, S. G. & Mackey, H. R. (2019). Greywater recycling in buildings using living walls and green roofs: A review of the applicability and challenges. *Science of The Total Environment*, 652, 330-344. <https://doi.org/10.1016/j.scitotenv.2018.10.226>

For Peer Review